

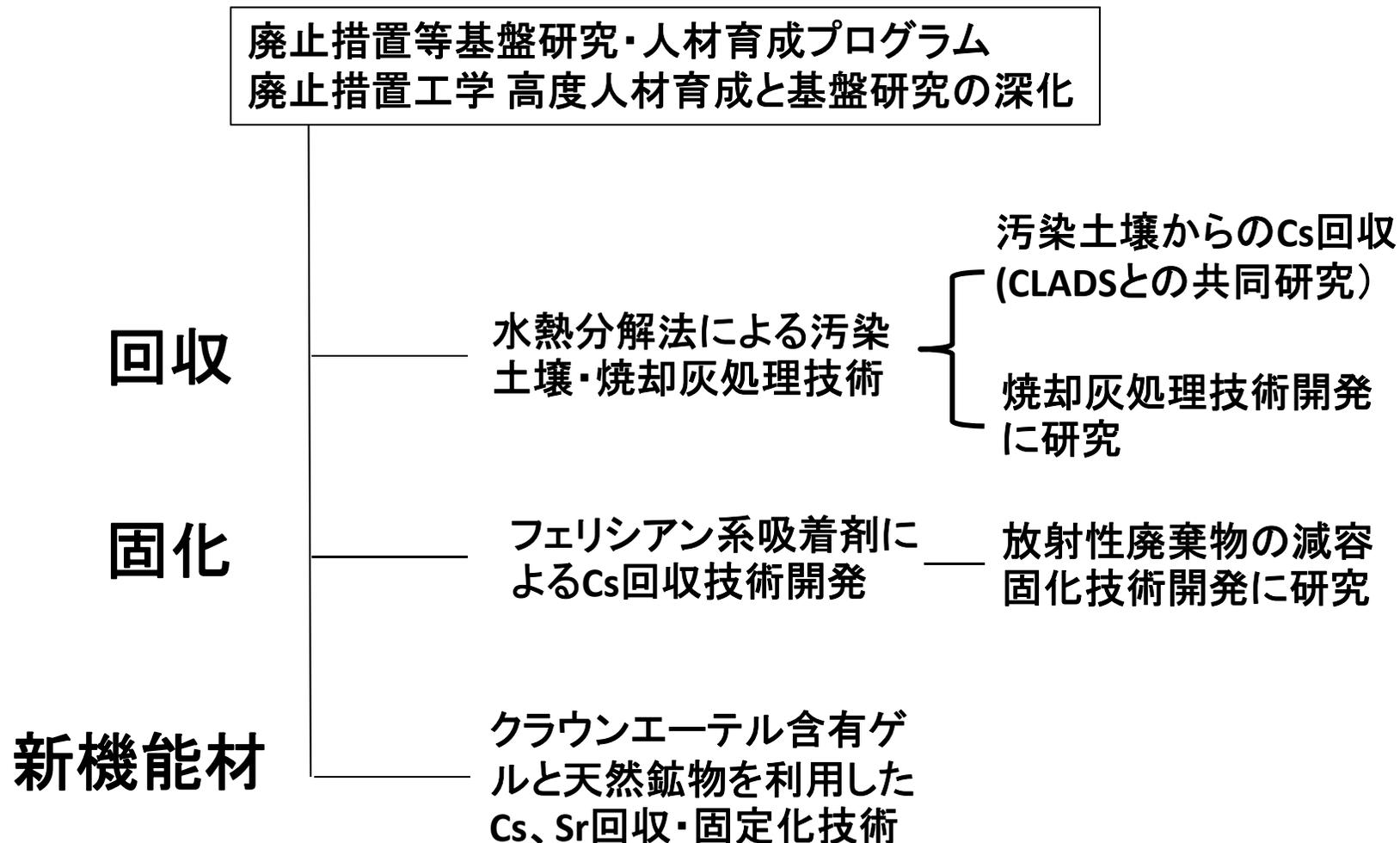
# 土壌の汚染除去と回収・固化技術-1

## 水熱分解法による汚染土壌・焼却灰処理技術

東京工業大学  
科学技術創成研究院  
先導原子力研究所  
竹下 健二

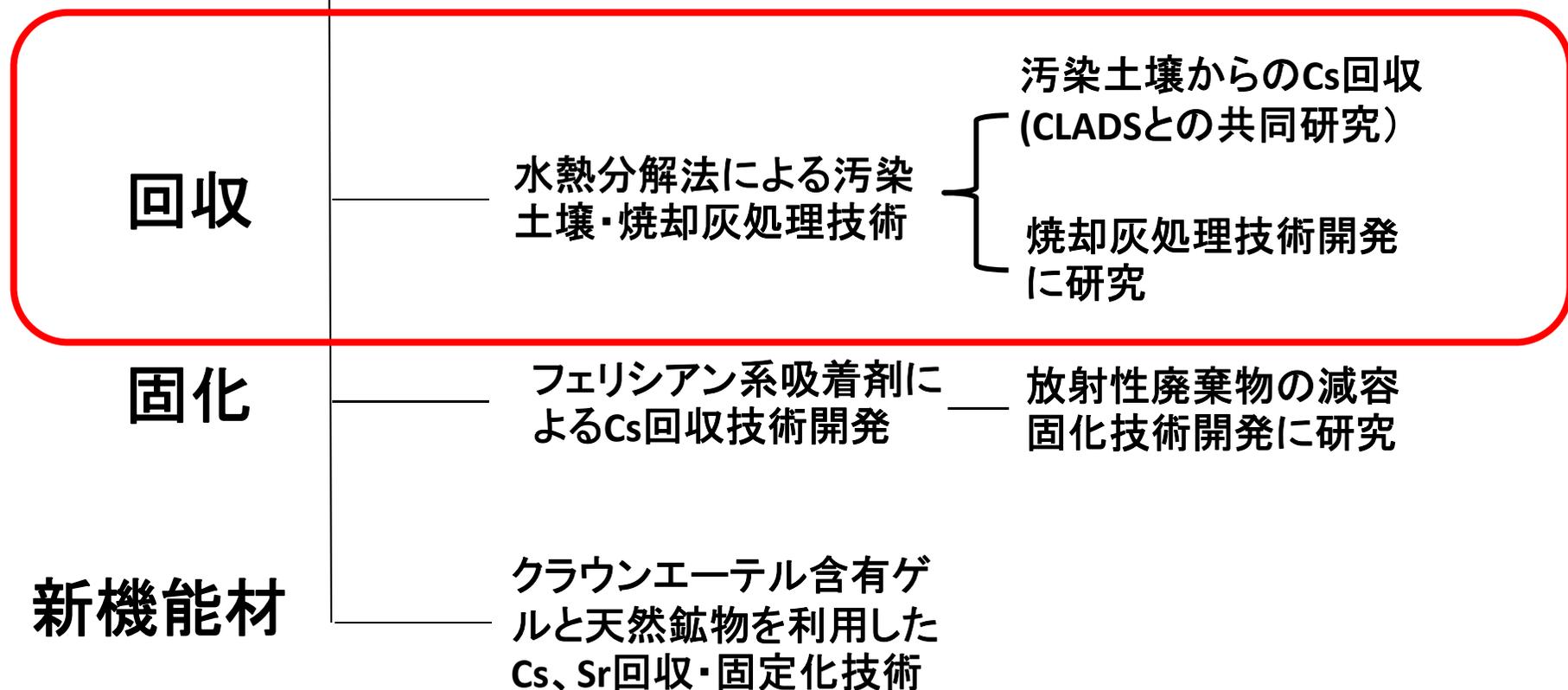
廃止措置技術・人材育成フォーラム  
2017年1月20日

# 現在進めている研究



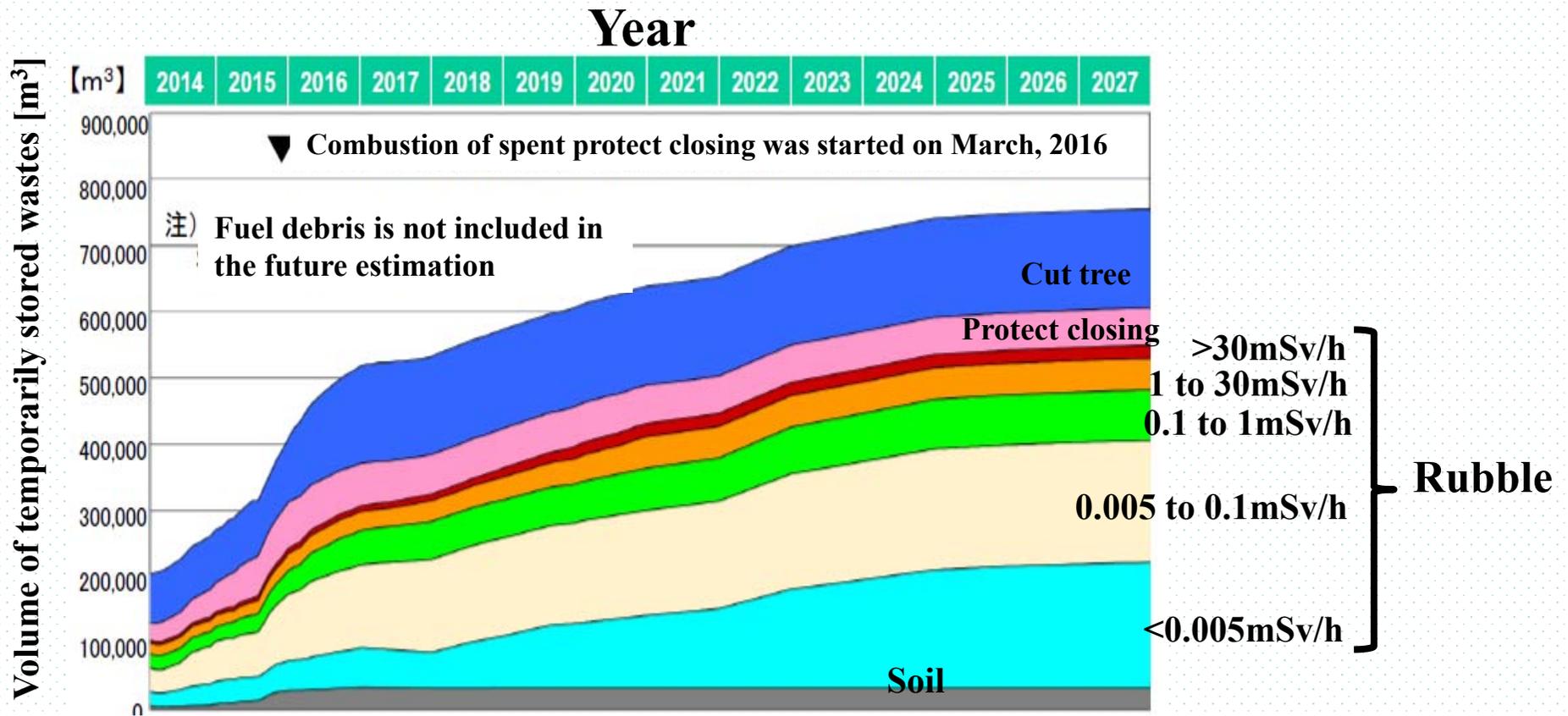
# 現在進めている研究

廃止措置等基盤研究・人材育成プログラム  
廃止措置工学 高度人材育成と基盤研究の深化



# Estimation of Solid Waste Generation

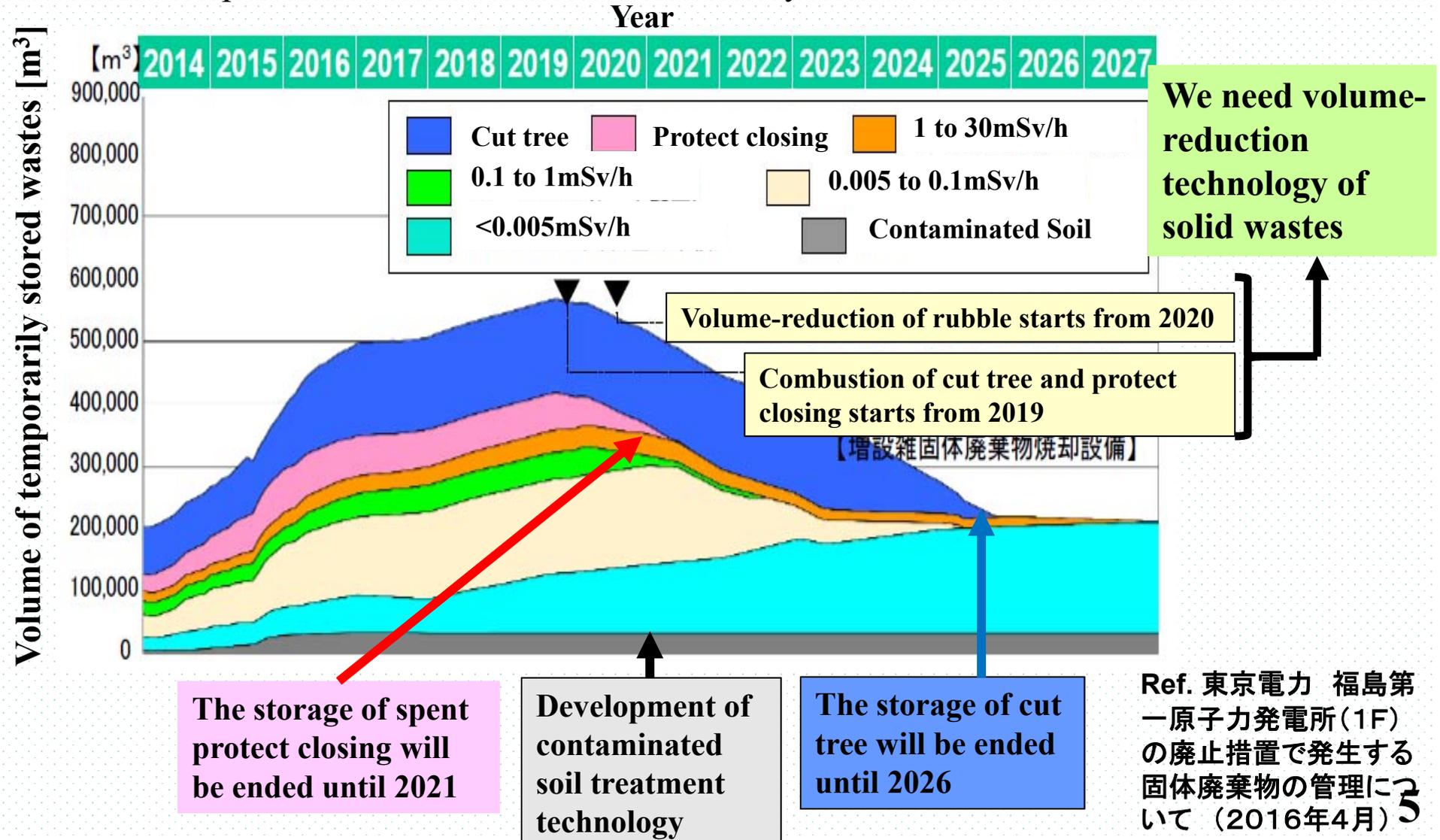
- This figure shows the time change of the volume of solid waste in temporary storage facility.
- If any measures to volume-reduce the solid wastes are not taken, total volume of solid wastes will be increased to **about 750,000 m<sup>3</sup>** by 2028.



\* Rubble means solid wastes including contaminated metal, concrete and soil

# Estimation of Solid Waste Generation with Volume-reduction Measures

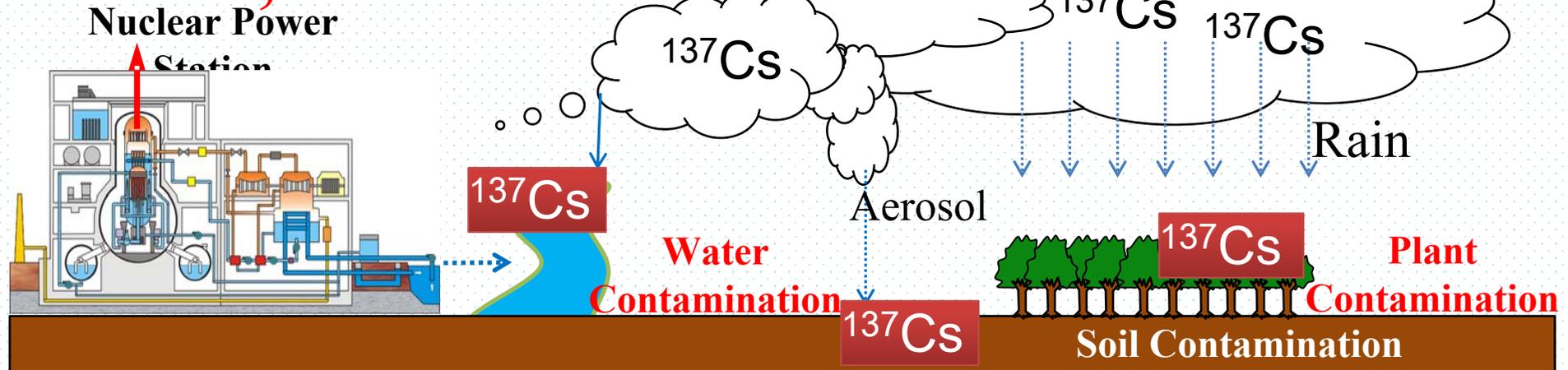
- Several volume-reduction measures such as combustion of cut tree and protective closing and the volume reduction of rubble, are considered. If these measures are attained, the volume of solid waste in temporary storage facility can be reduced. The temporary storage of solid wastes except low-radiation rubble will be ended by 2026.



# Release of Radioactive Nuclides to Environment

## Release to atmosphere

### Cs-134, Cs-137

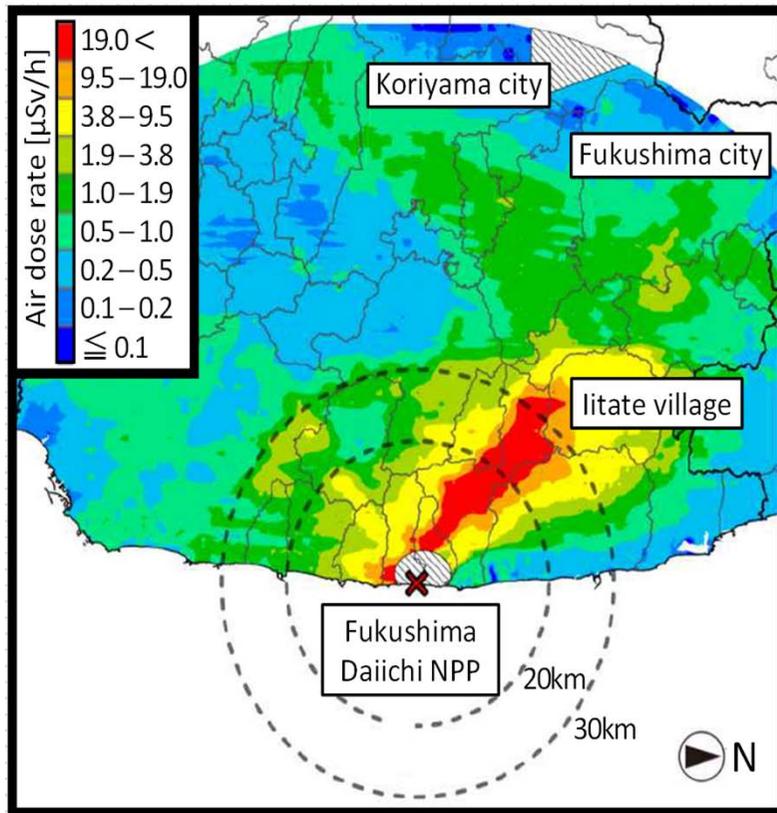


- Cs and I were released to the atmosphere by the explosion of reactor buildings and fallouts of radioactive elements contaminated widely in Eastern Japan.
- A short-lived nuclide, I-131, has already decayed. Nowadays Cs-134 and Cs-137 are main radioactive nuclides of environmental pollution.

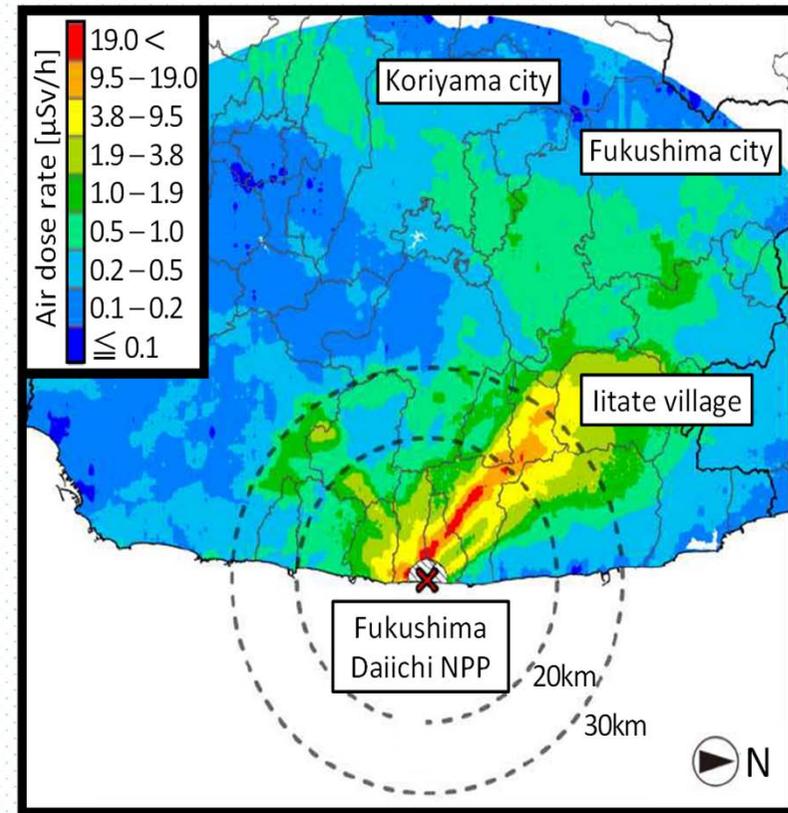
Total amount of released  $^{137}\text{Cs}$  : 15 PBq  $^{134}\text{Cs}$  : 18 PBq  
Data By Nuclear and Industrial Safety Agency (2011)

1PBq=10<sup>15</sup>Bq

# Circumstance of Cs deposition in Eastern Japan



**April 29, 2011**



**September 28, 2013**

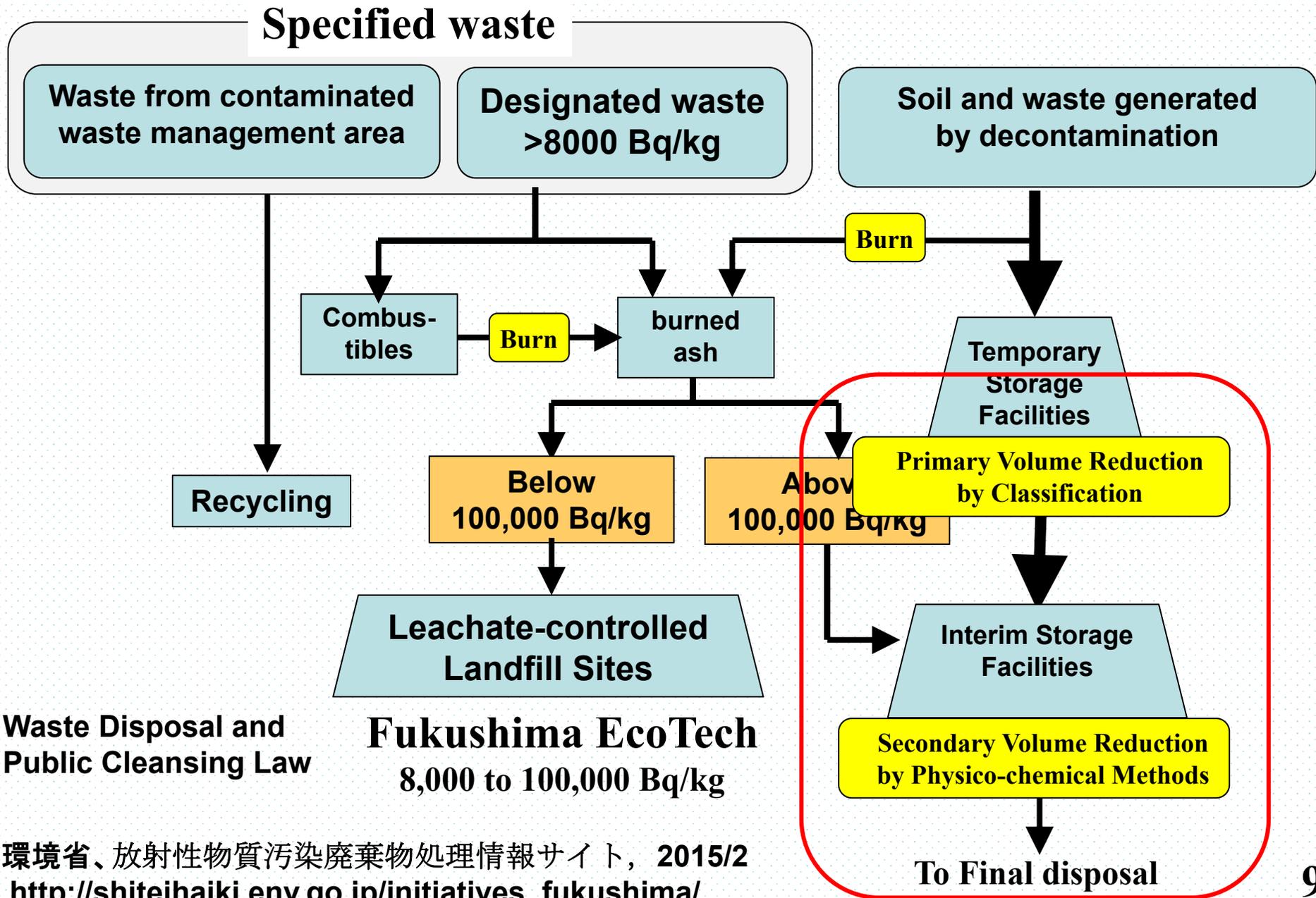
The width of plume becomes narrower as time elapses. Radioactivity of  $^{137}\text{Cs}$  and  $^{134}\text{Cs}$  is reduced by not only the physical decay, but also the weathering effect such as wind and rain. The decreasing rate of the radioactivity is faster than that by the physical decay.

# Classification of Contaminated Waste (2011)

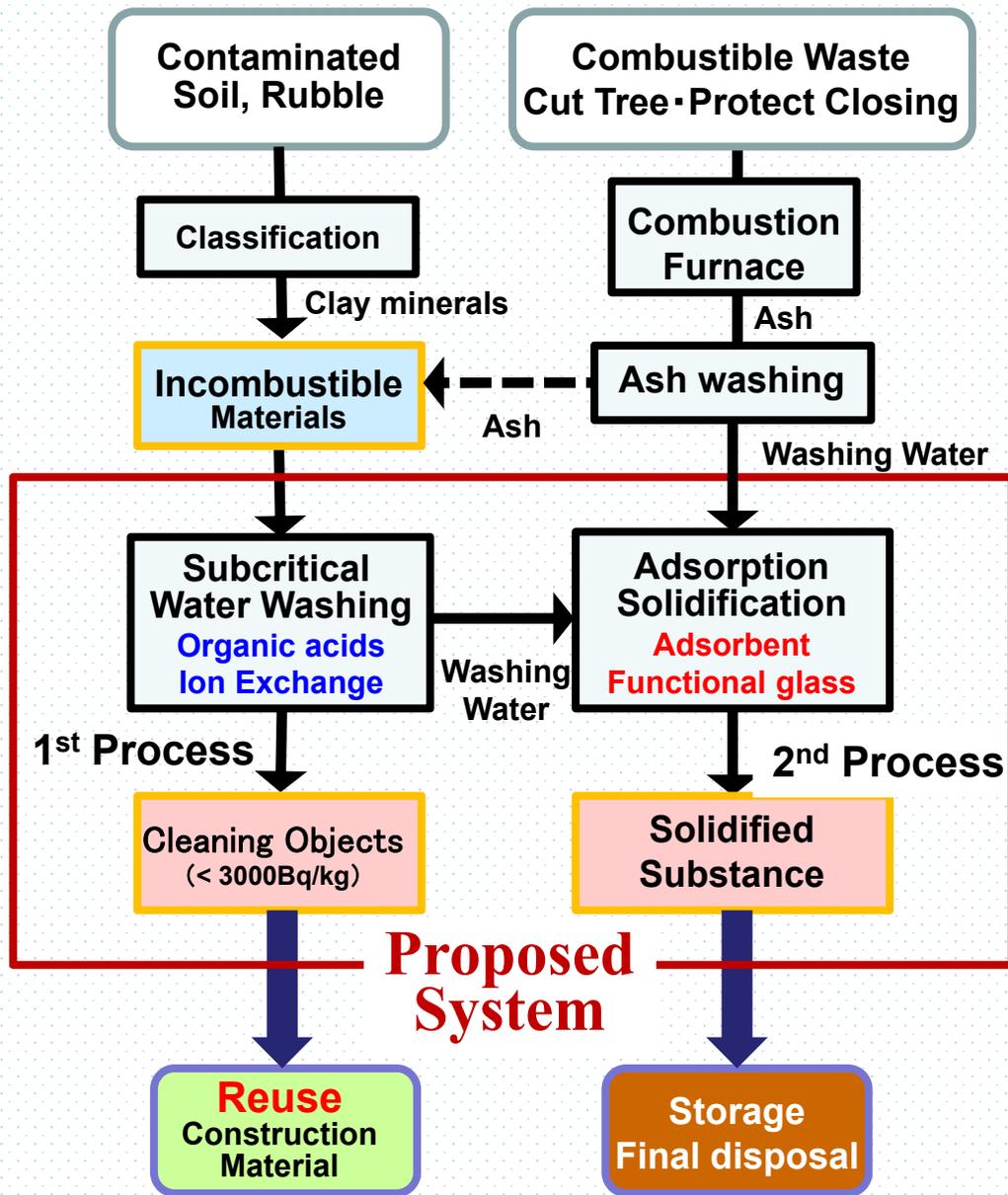
	Fukushima Pref.			Outside Fukushima Pref.	
Classification of Waste	Specified Waste		Soil and waste generated by decontamination	Specified Waste	Soil and waste generated by decontamination
	Waste from Contaminated waste management area (Wastes from restricted areas and deliberate evacuation areas)	Designated Wastes (Paddy straw and wastes generated from water and sewage facilities and waste incineration facility, whose radiation doses are above 8000 Bq/kg.)		Only designated waste	
Weight of generated waste	about 500,000 ton	about 60,000 ton/yr	about 15 to 31 million m <sup>3</sup>	about 80,000 ton/yr	about 1.4 to 13 million m <sup>3</sup>

Contaminated waste is classified as three categories, waste from contaminated waste management area, designated waste (> 8000 Bq/kg) and soil and waste generated by decontamination. In 2011, Ministry of the Environment estimated weights of the wastes. These values are very large.

# Flow Diagram of Specified Waste and Contaminated Soil



# Proposal of Solid Waste Treatment System



We propose a new solid waste treatment system with two processes to realize the concept of solid waste treatment.

## 1<sup>st</sup> Process (HTT process)

① Radioactive elements (Cs) in solid waste are recovered in **subcritical water with organic acids or alkaline and alkaline earth ions**.

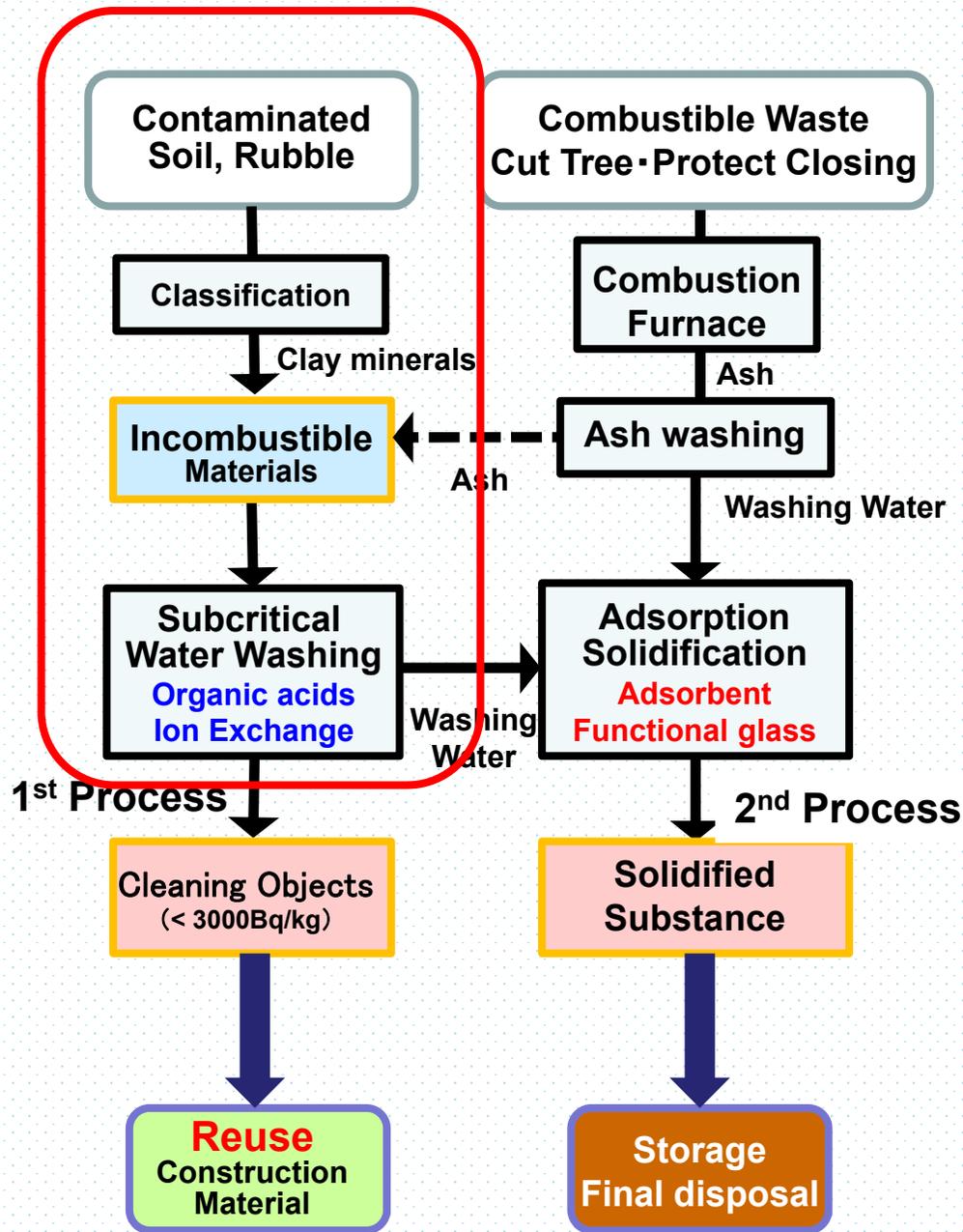
② Cleaning objects can be reused industrially (e.g. construction material)

## 2<sup>nd</sup> Process

① Radioactive elements (Cs) in washing water are concentrated in **highly selective adsorbent or functional glass**.

② Radioactive elements (Cs) are solidified in stable inorganic materials (e.g. glass, heteropoly acid). A small amount of radioactive waste is stored finally.

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# Cs Adsorption in Various Clays

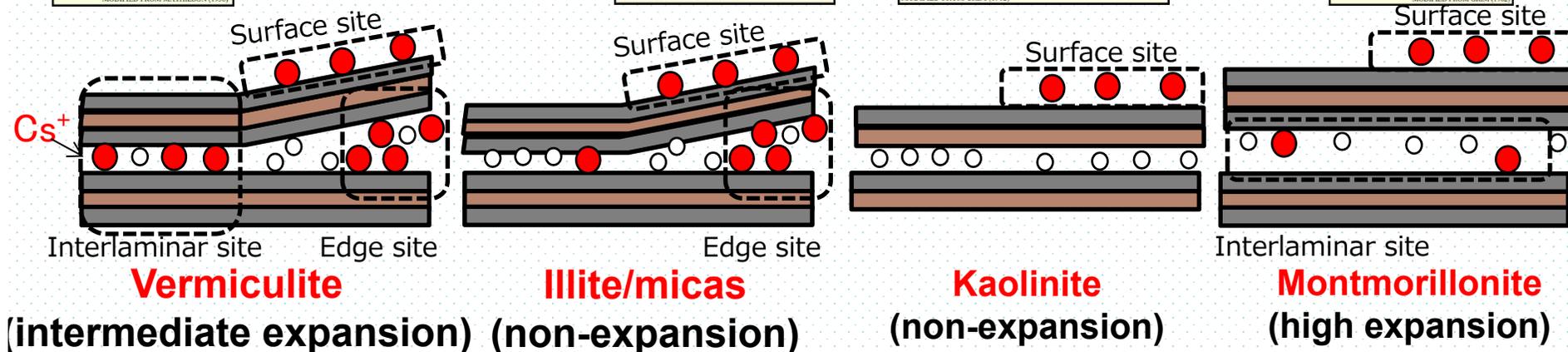
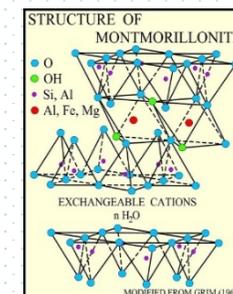
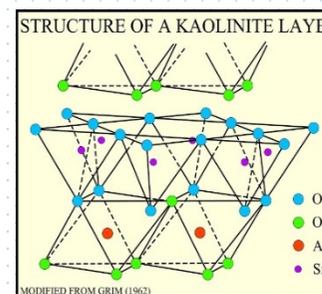
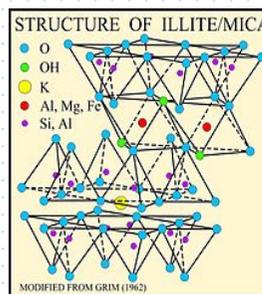
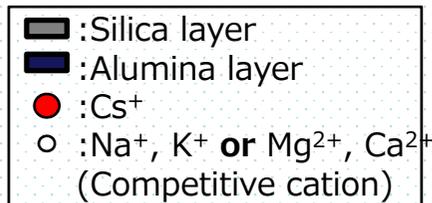
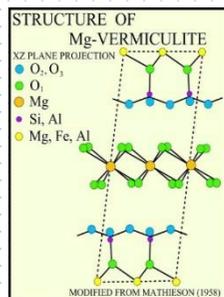
## Clay Minerals

Montmorillonite: 2:1 type, high expansion

Vermiculite: 2:1 type, intermediate expansion

Illite: 2:1 type, non-expansion

Kaolinite: 1:1 type, non-expansion

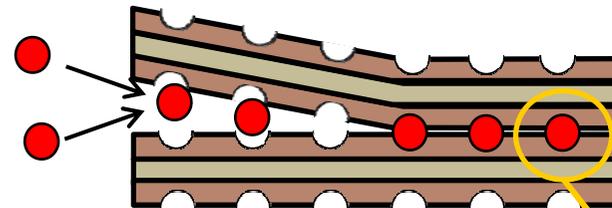


# Adsorption of Cs(I) into Clay Mineral

Contaminated soil with Cs-137 ( $t_{1/2}=30\text{yr}$ )

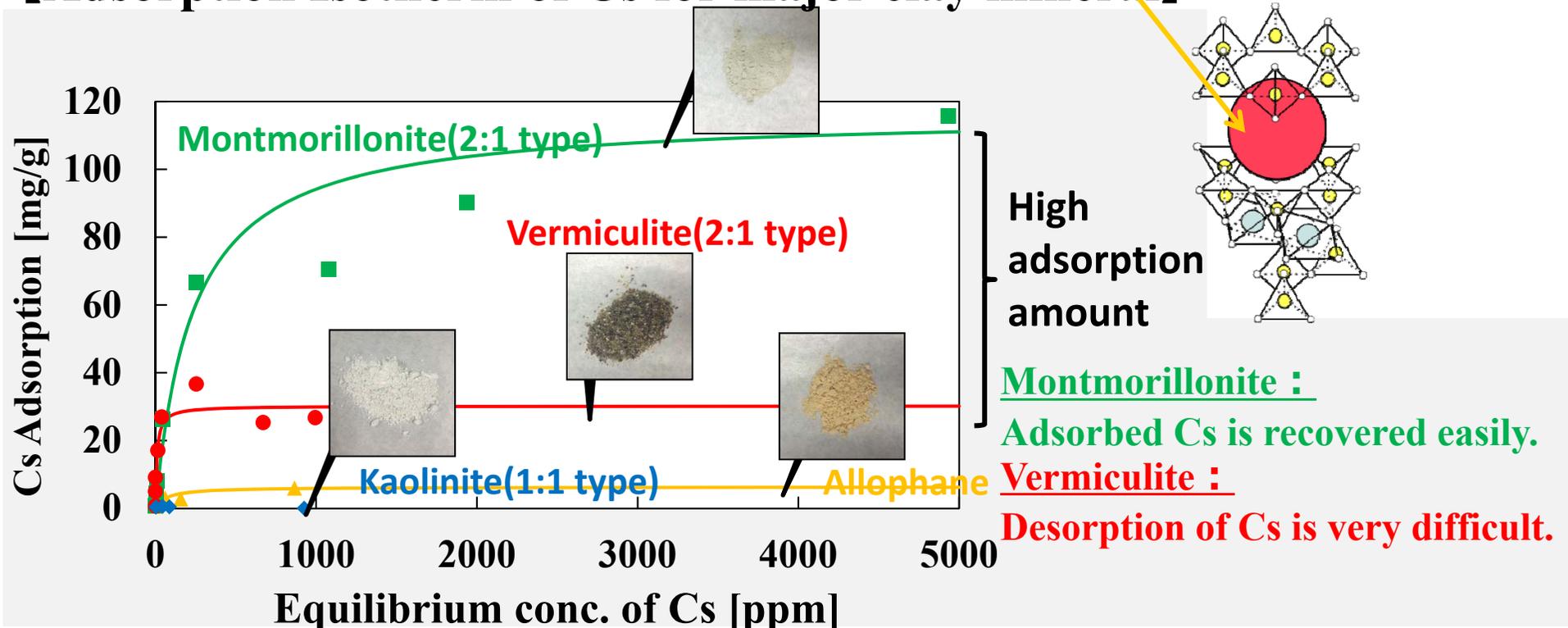
2:1 type layered clay mineral

Cs is adsorbed strongly into 2:1 type layered clay mineral of soil



	:tetrahedral layer (Si, Al)
	:octahedral layer (Al, Mg, Fe)
	:Cs <sup>+</sup>

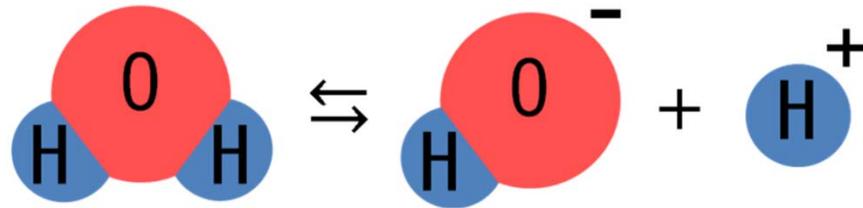
**[Adsorption Isotherm of Cs for major clay mineral]**



If Cs is desorbed from vermiculite, the recovery of Cs from soil is improved

# Subcritical Water

Ionic product of water becomes larger in subcritical condition under high temperature (about 200°C) and high pressure (about 3MPa).



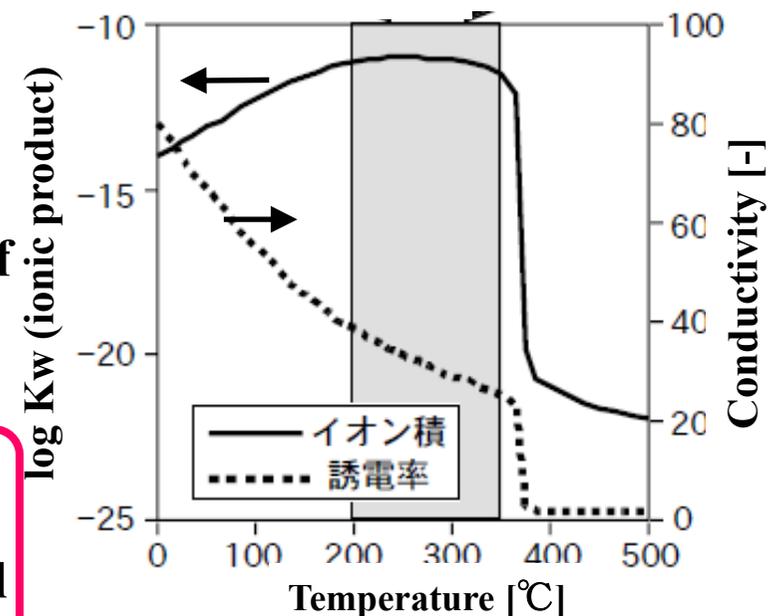
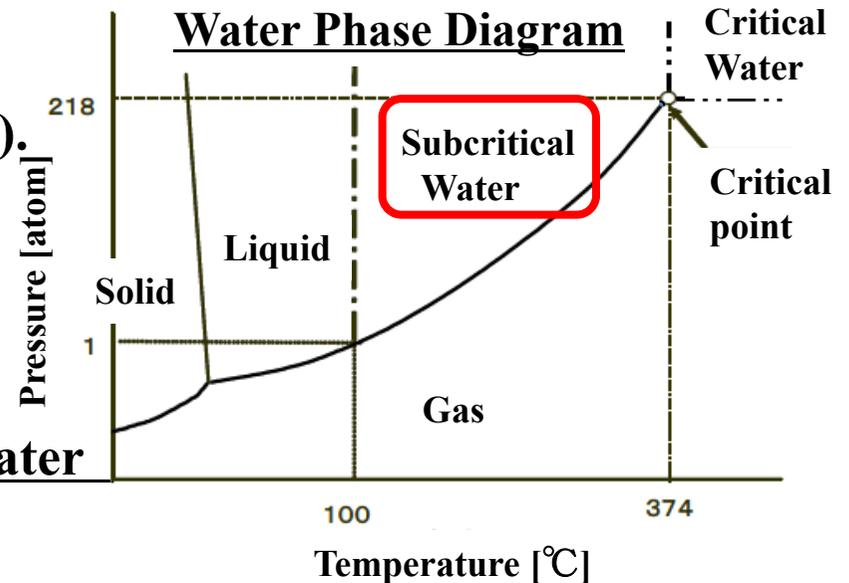
## Specific Chemical Properties of Subcritical Water

### ① High-speed Hydrolysis Effect

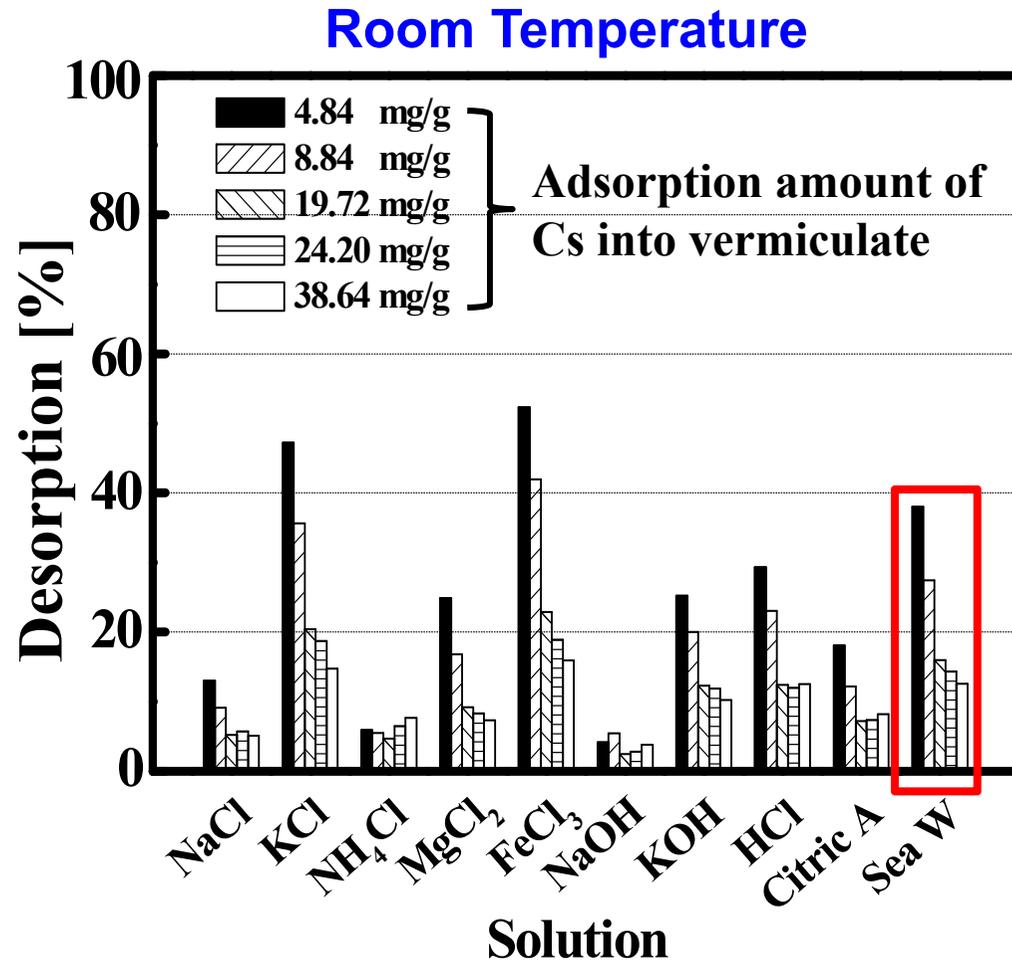
- Cs adsorbed in organic materials can be recovered to water phase by the decomposition of organic materials.
- Cs adsorbed in inorganic materials is also recovered by the high-speed decomposition of inorganic materials by subcritical water with organic acids.

### ② High-speed Ion Exchange Effect

Cs adsorbed in inorganic materials can be recovered to the aqueous phase by subcritical water with alkaline and alkaline earth ions.



# Ion-exchange Recovery of Cs from Vermiculate at Room Temp.

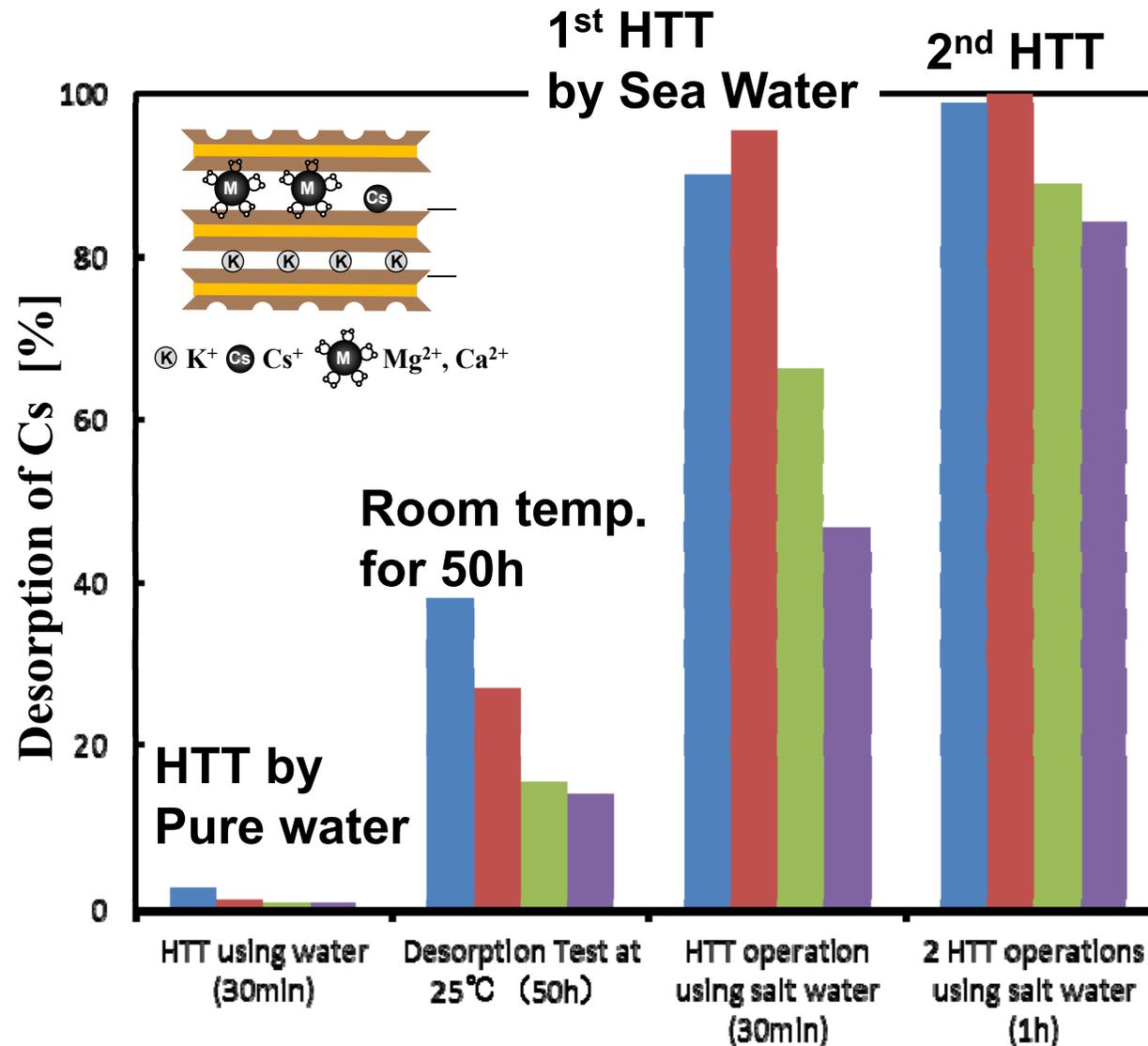


## Experimental Conditions

Volume of salt solution = 10ml  
 Weight of Vermiculate = 100mg  
 Conc. of salt = 0.1M  
 Temperature = 25 °C  
 Operation time = 24h.

- Ability:  $Fe^{3+} > K^+ > S.W > H^+ > Mg^{2+} > Citric\ Acid > Na^+ > NH_4^+$  .
- Sea water has a relative high ion exchange ability.
- Ion-exchange technique is useful for the recovery of Cs from Vermiculate.

# Ion Exchange of Cs Adsorbed into Vermiculite



## Experimental Conditions

Volume of solution=50ml  
 Weight of Vermiculite =0.5g  
 Temperature =25°C  
 Operation Time= 30 min  
 Conc. of salt -0.5 M

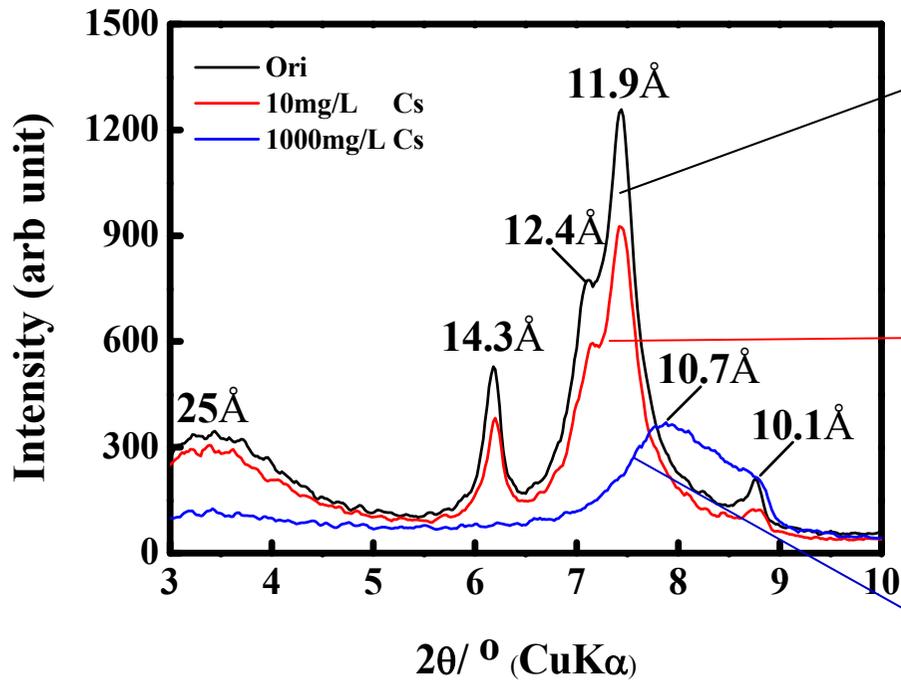
Adsorption amount of Cs into Vermiculite

- 4.84 mg/g
- 8.84 mg/g
- 19.7 mg/g
- 24.2 mg/g

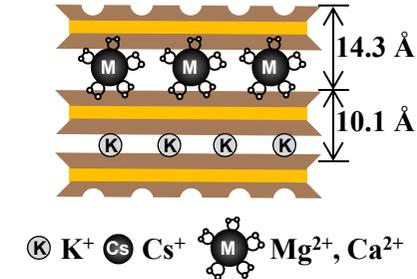
Conc. of  $^{133}\text{Cs}$  in soil = 10mg/kg

- HTT using water is not useful for the desorption of Cs
- Desorption rate of Cs with salt solution at 25°C is very slow.
- Sea water+HTT promotes the desorption of Cs.

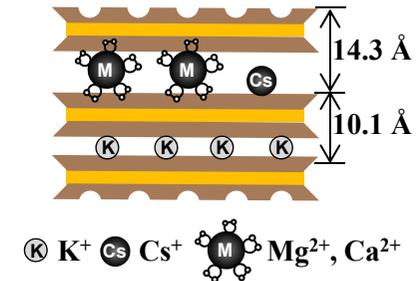
# XRD Analysis of Cs-Vermiculite



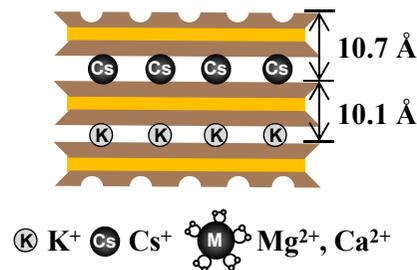
## Original sample



## Low loading sample



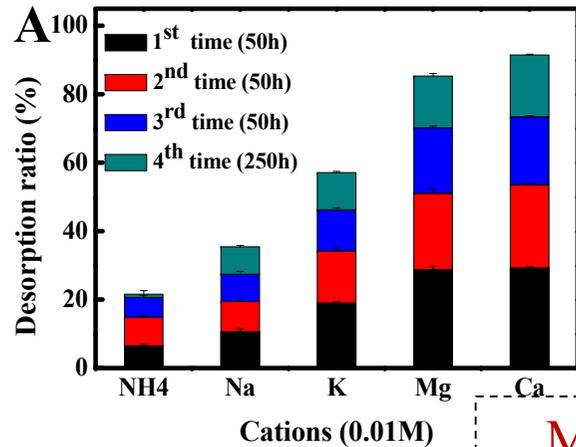
## High loading sample



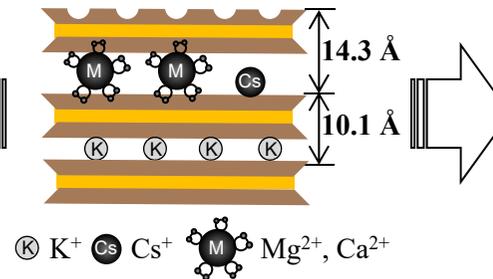
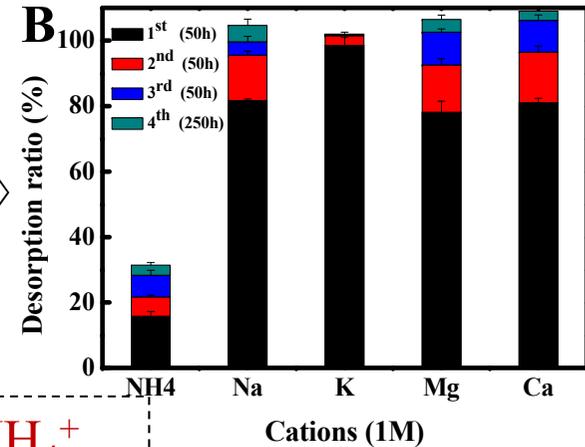
- ◆ 25 Å: Regular interstratification of Mg and K layer
- ◆ 14.3 Å: Hydrated  $\text{Mg}^{2+}$  layer
- ◆ 12.3 Å: Partial hydrated  $\text{Mg}^{2+}$  layer
- ◆ 11.9 Å: Random interstratification of Mg and K layer
- ◆ 10.7 Å: Cs collapsed layer
- ◆ 10.1 Å: Biotite  $\text{K}^+$  layer

# Cs Desorption from Vermiculite with low Cs loading (25°C)

Low ionic strength (0.01M)



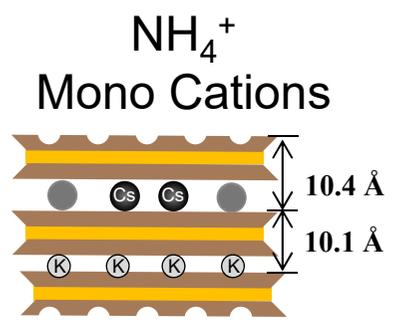
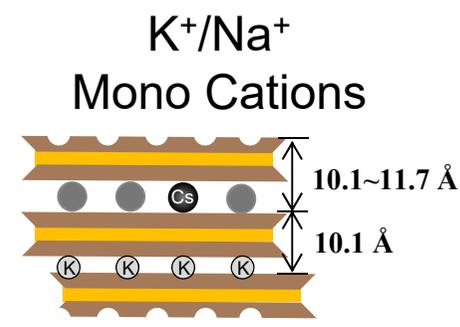
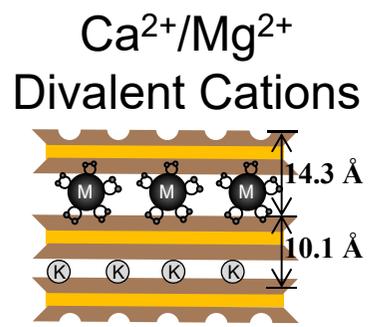
High ionic strength (1M)



{ Hydrous radius:  $Mg(Ca) > Na > K(NH_4)$   
 { Hydration energy:  $Mg(Ca) > Na > K(NH_4)$   
 { Ion valence:  $Mg(Ca) > Na = K(NH_4)$

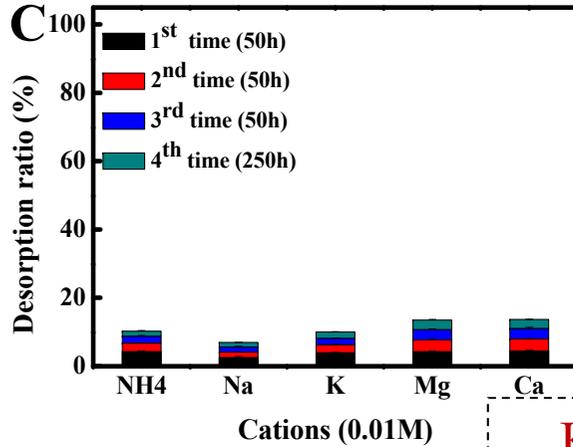
- Cs is desorbed more easily by divalent cation.
- High concentrated  $K^+(Na^+)$  desorbs Cs easily.
- $NH_4^+$  desorb the least  $Cs^+$ .

After Cs desorption

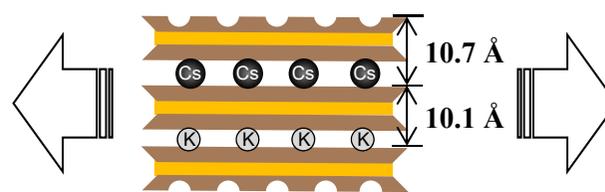
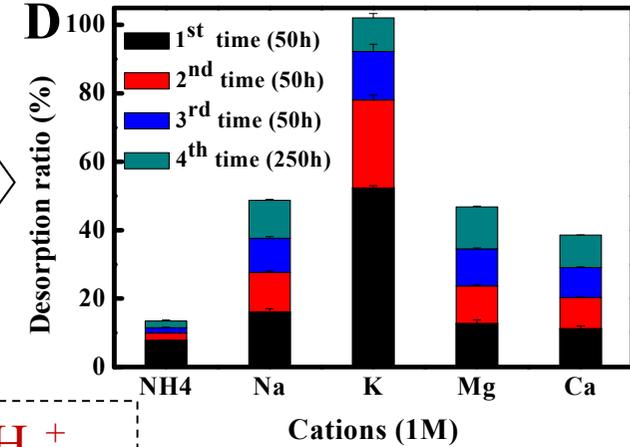


# Cs Desorption from Vermiculite with High Cs Loading (25°C)

Low ionic strength (0.01M)



High ionic strength (1M)

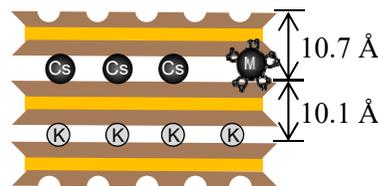


Radius:  $Mg(Ca) > Na > K(NH_4)$   
 Energy:  $Mg(Ca) > Na > K(NH_4)$   
 Valence:  $Mg(Ca) > Na = K(NH_4)$

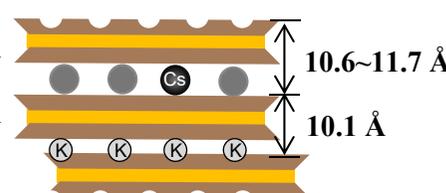
- Cs is rarely desorbed by each cation in dilute solution.
- Cs is desorbed more easily by high concentration of  $K^+$ .
- High concentrations of  $Na^+$ , and  $Mg^{2+}/Ca^{2+}$  are useful for desorption.
- $Cs^+$  is rarely desorbed by  $NH_4^+$ .

After Cs desorption

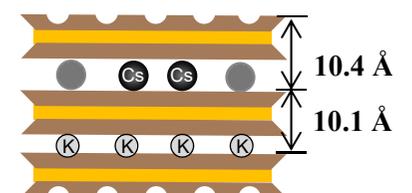
$Ca^{2+}/Mg^{2+}$   
Divalent Cations



$K^+/Na^+$   
Mono Cations

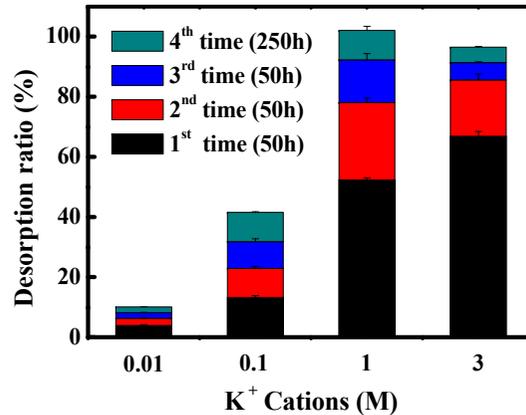


$NH_4^+$   
Mono Cations

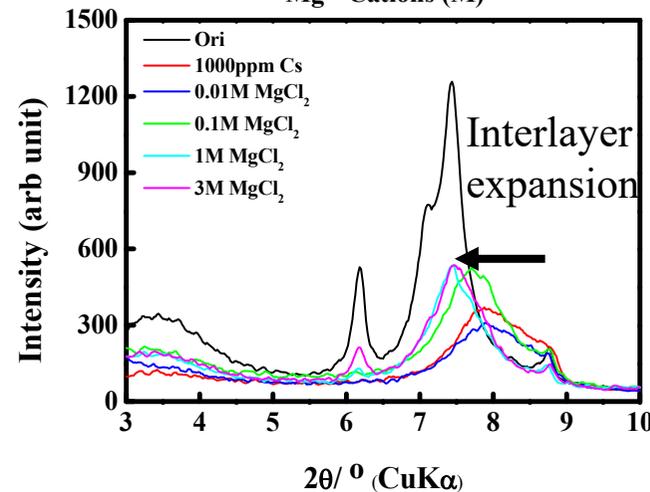
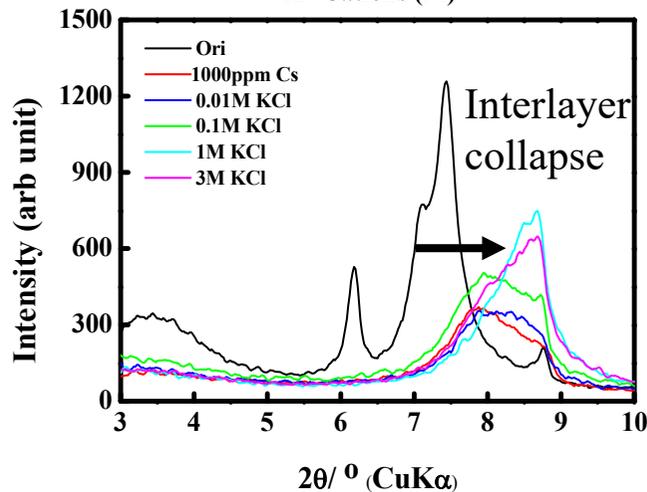
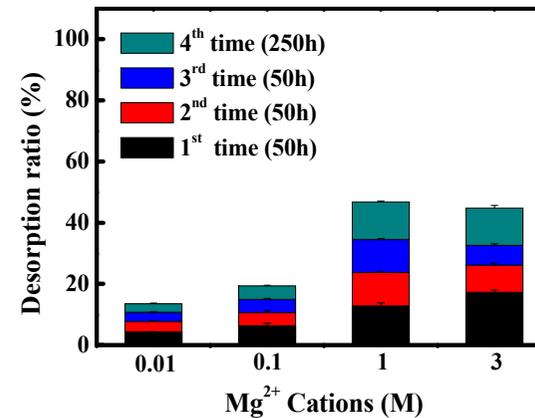


# Cs Desorption from Vermiculite with High Cs Loading (25°C)

## Effect of K<sup>+</sup> conc.

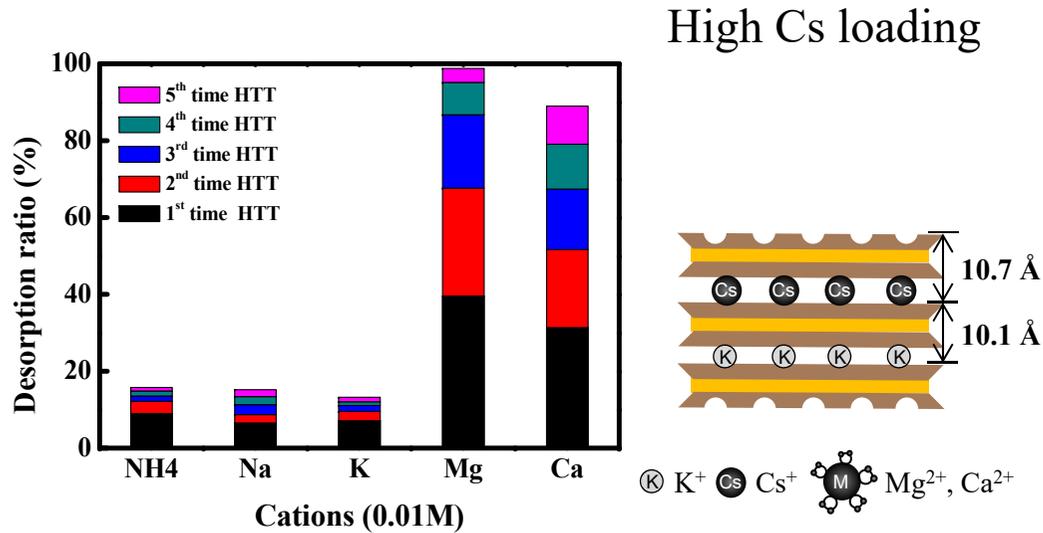


## Effect of Mg<sup>2+</sup> conc.



- Cs desorption is promoted by the high concentration of cation.
- Diffusion ability of divalent cations to the collapsed interlayer is very weak at room temperature due to their larger ionic radius.

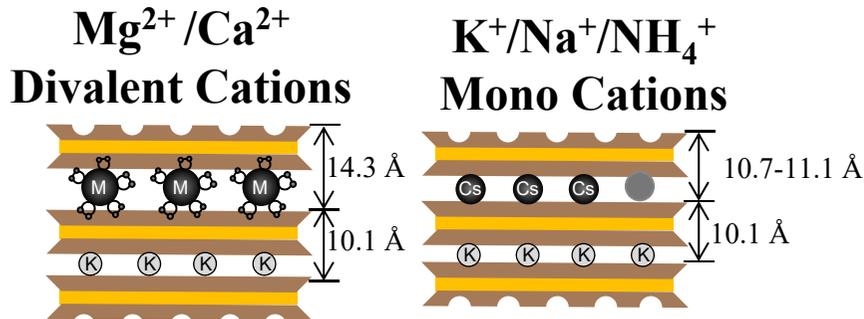
# Cs Desorption from Vermiculite with High Cs Loading (250°C)



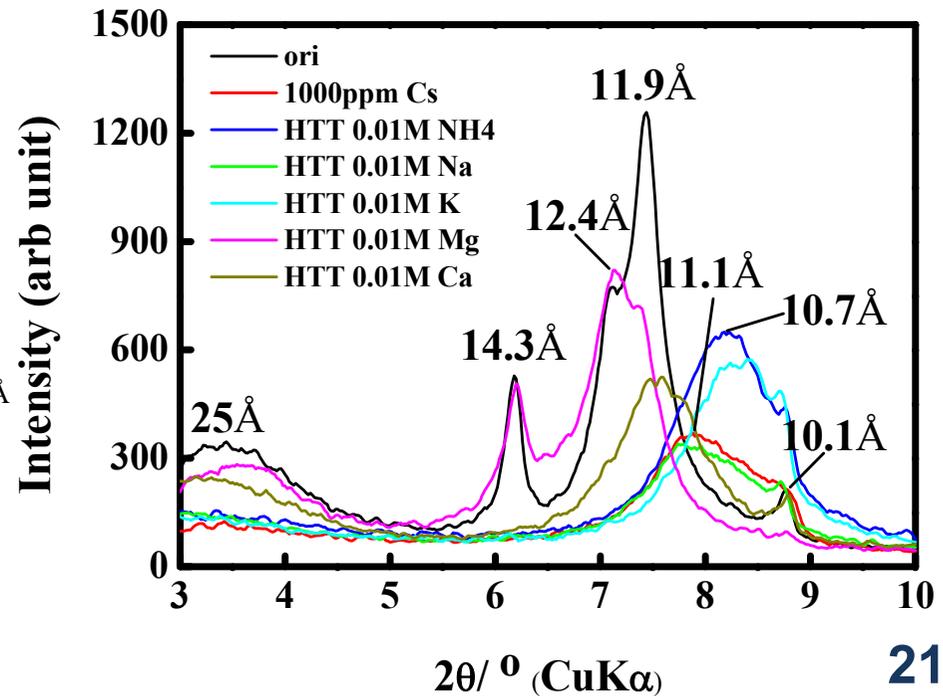
- The results are contrary to those at 25 °C.
- The Cs desorption is promoted by the subcritical water washing with divalent cations.
- The use of monovalent cations is not useful for the desorption of Cs.
- Both Cs and original K are desorbed by Mg<sup>2+</sup>.



After Cs desorption

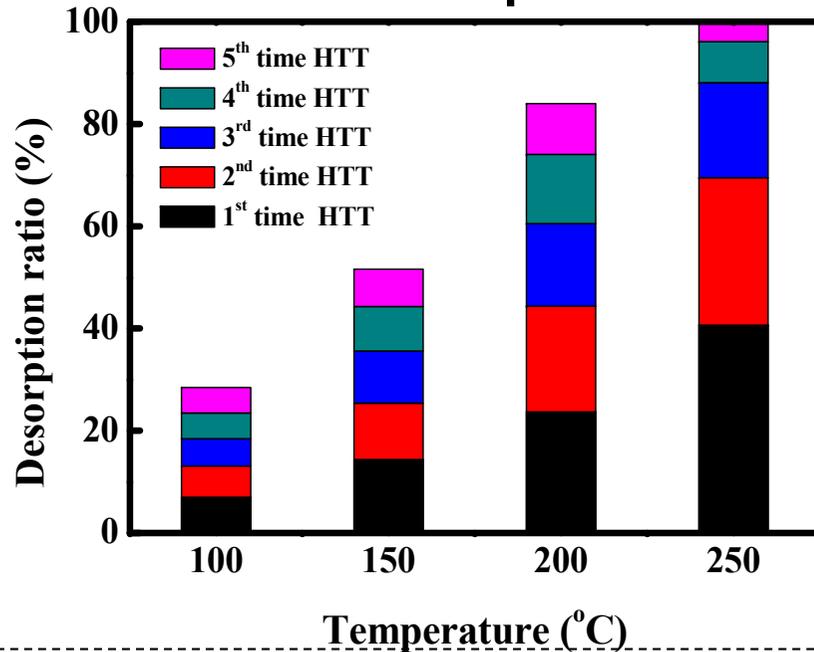


## XRD analysis



# Cs Desorption from Vermiculite with High Cs Loading (250°C)

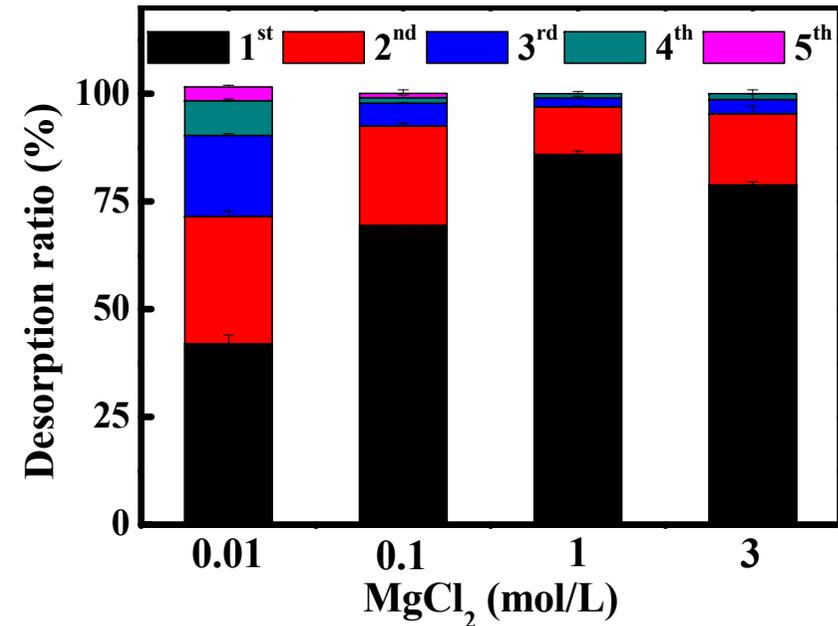
## Effect of temperature



(100-250°C), S/L=0.5g/50ml

- Mg<sup>2+</sup>濃度0.01M、L/S=50ml/0.5g
- 亜臨界水温度の増加→高速イオン交換

## Effect of ionic strength

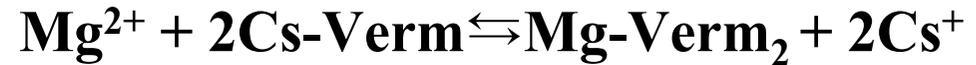
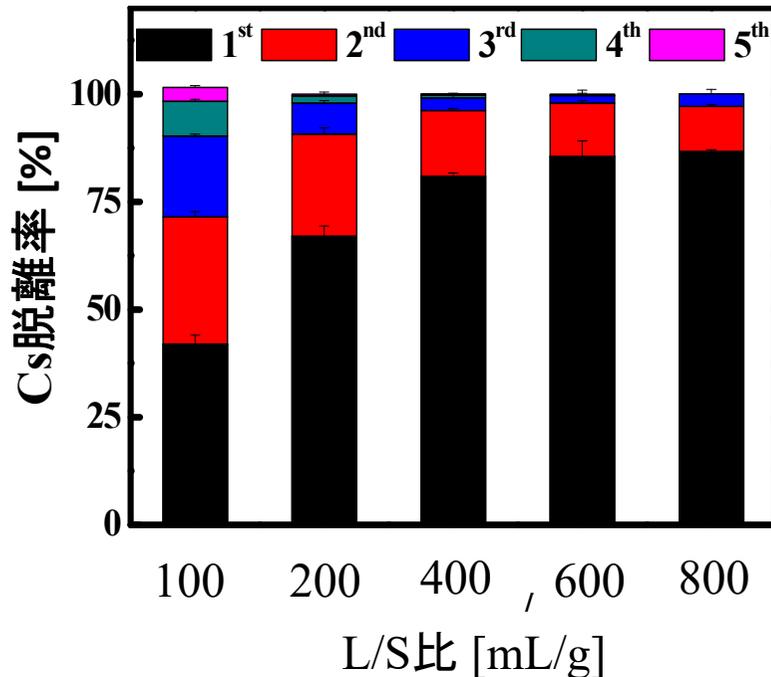


HTT(250°C), S/L=0.5g/50ml

- Mg<sup>2+</sup>濃度の増加に従ってイオン交換速度が増加し、Cs脱離率増加
- 1M程度で最大。高濃度にし過ぎるとむしろCs脱離率低下。
- 1M程度の濃度では1回の洗浄でCs脱離率80%程度が期待できる

# Cs Desorption from Vermiculite with High Cs Loading (250°C)

Effect of L/S ratio



Enhanced Cs Desorption

1.  $\text{Mg}^{2+}$  concentration  $\uparrow$

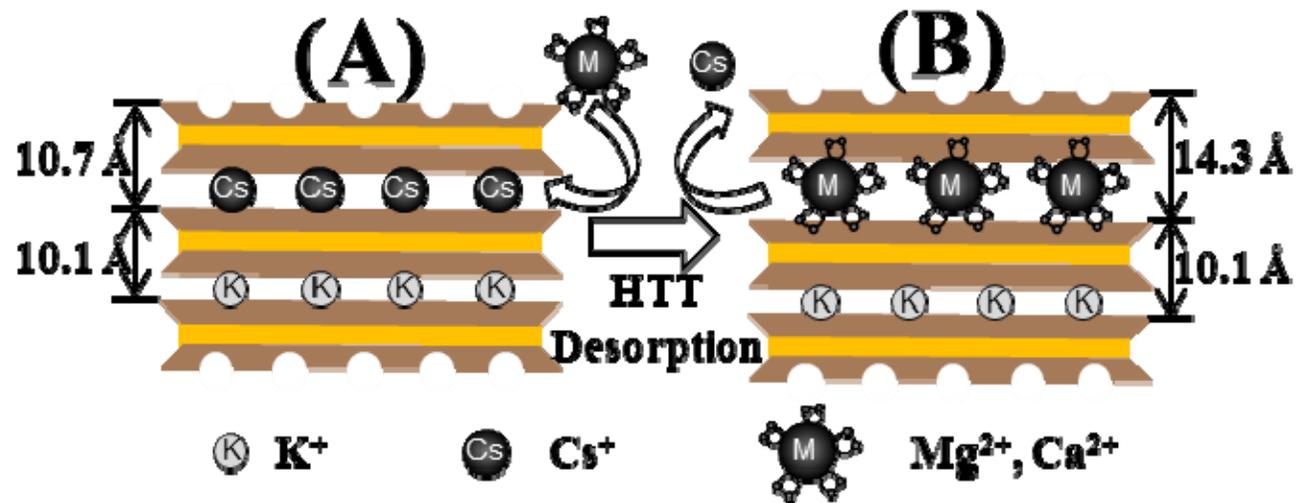
2.  $\text{Cs}^+$  concentration  $\downarrow$

3. Temperature  $\uparrow$

HTT (250°C),  $\text{Mg}^{2+}=0.01\text{M}$

- L/S比が増加するに従ってイオン交換速度が増加し、Cs脱離率増加
- 0.01Mの低イオン強度においても、L/S比が大きくなると、2回程度の洗浄で100%のCs脱離が可能になる。

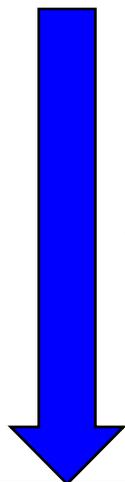
# Cs Desorption from Vermiculite with High Cs Loading (250°C)



# 採取土壌について

三菱マテリアル那珂エネルギー開発研究所内(茨城県那珂市)にて土壌を採取

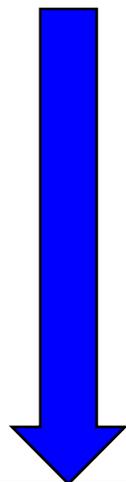
雨樋の排水口周辺など、  
雨水により環境中の放射性セシウムの  
流入が見込まれる場所から採取



汚染  
土壌

採取した土壌を2mm未満に  
分級後、混合

- ・約3kg所有
- ・混合した土壌の放射能濃度  
 $1.55 \times 10^4$  Bq/kg



非汚染  
土壌

採取した土壌を2mm未満に  
分級後、混合

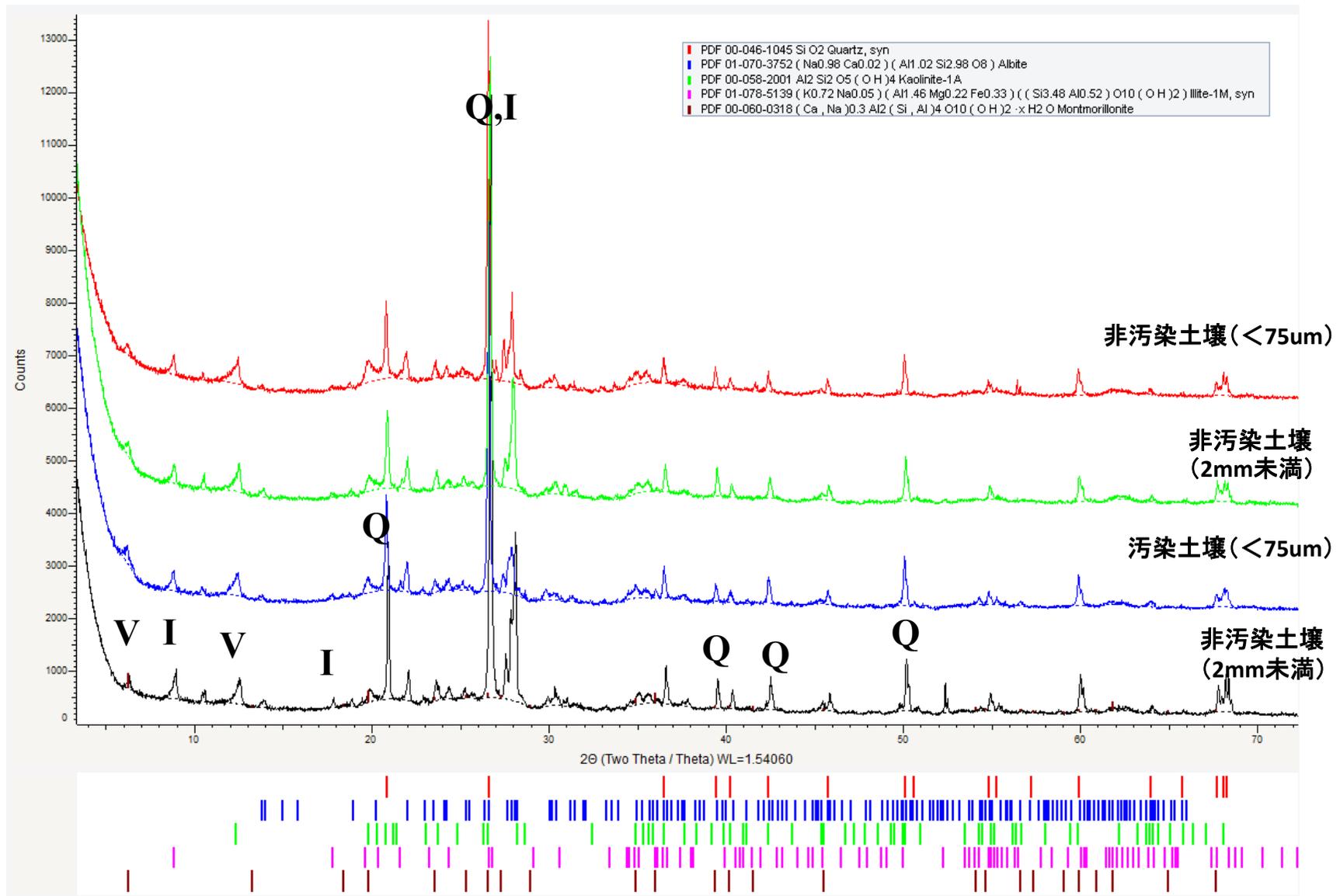
- ・約8kg所有
- ・混合した土壌の放射能濃度  
 $2.12 \times 10^2$  Bq/kg

細粒分を念頭にさらに<75 $\mu$ mに分級



土壌 (<75 $\mu$ m未満)

# 汚染土壌と非汚染土壌のXRD分析結果(<75 $\mu$ m分級前後の比較)



分級後土壌(<75 $\mu$ m)においても鉱物組成に大きな差異はない。

## (2) イオン含有亜臨界水によるCs脱離試験(ホット)

・実施条件例:

試験回数	対象物	添加イオン または添加物	添加イオン 濃度[M]	液固比 [ml/g]	保持温度 [°C]
全3回程度	ホット試験用試料 (汚染土壌の細粒物)	Mg <sup>2+</sup> or Mg <sup>2+</sup> + 0.3M Citric Acid	0.01 ~ 0.5	5 ~ 100	250

※コールド試験の結果から設定する。試験結果によっては同条件による繰り返し再現性試験をすることもある。

### 試験方法:

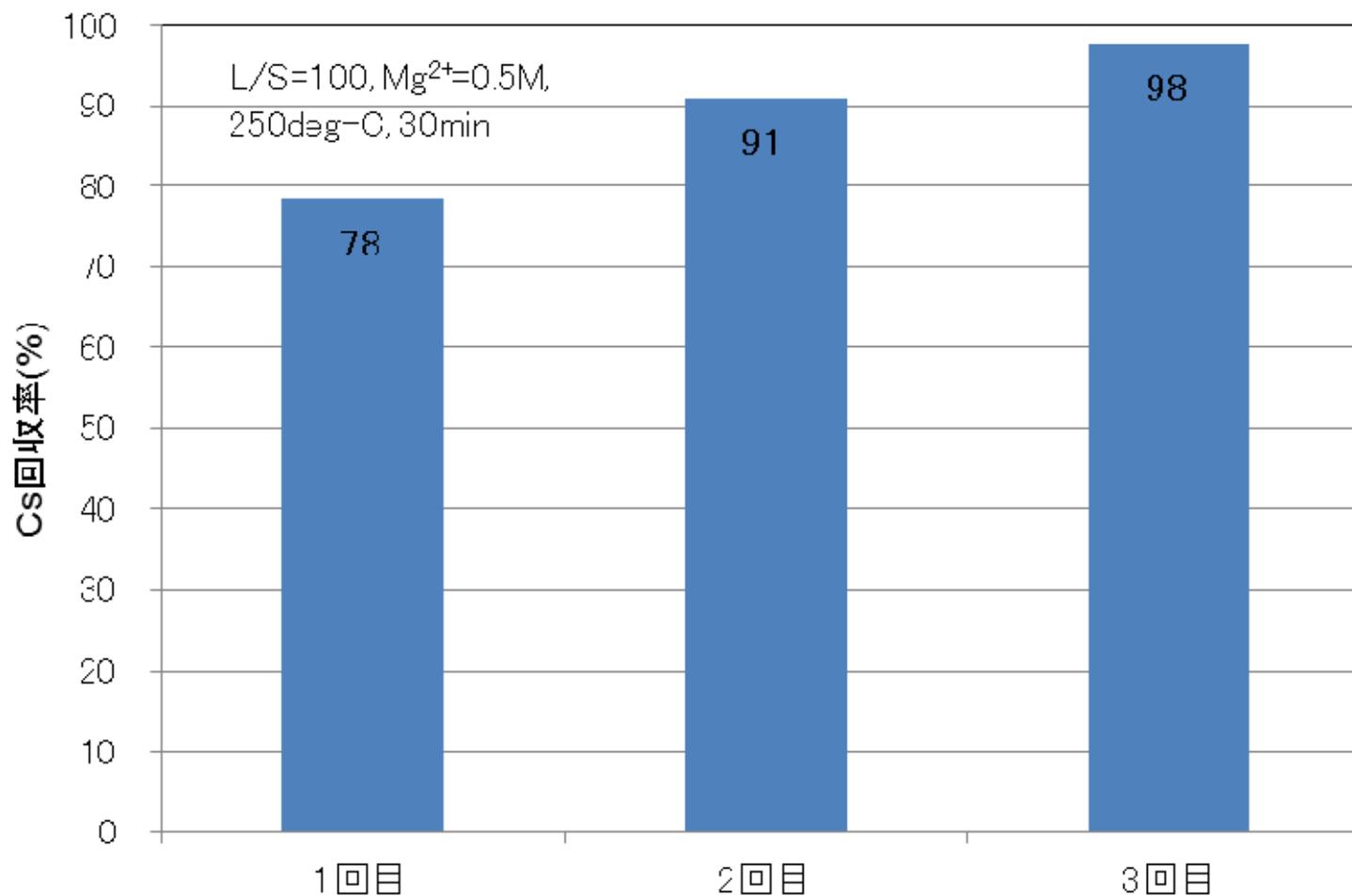
- コールド試験の結果から得られたCs脱離の最適条件(添加イオン組成、濃度、液固比、保持温度)を中心に、ホット試験用試料を用いて小型水熱処理装置で脱離試験を行う。
- 試験後の試料とイオン水の放射エネルギーをGe半導体検出器で測定し、Cs脱離率を求める。
- さらにコールド試験と同様に繰り返し試験を行い、Cs脱離率の変化を調べる。

### 評価項目:

- ・Cs脱離率、及び物質収支の評価を行う。

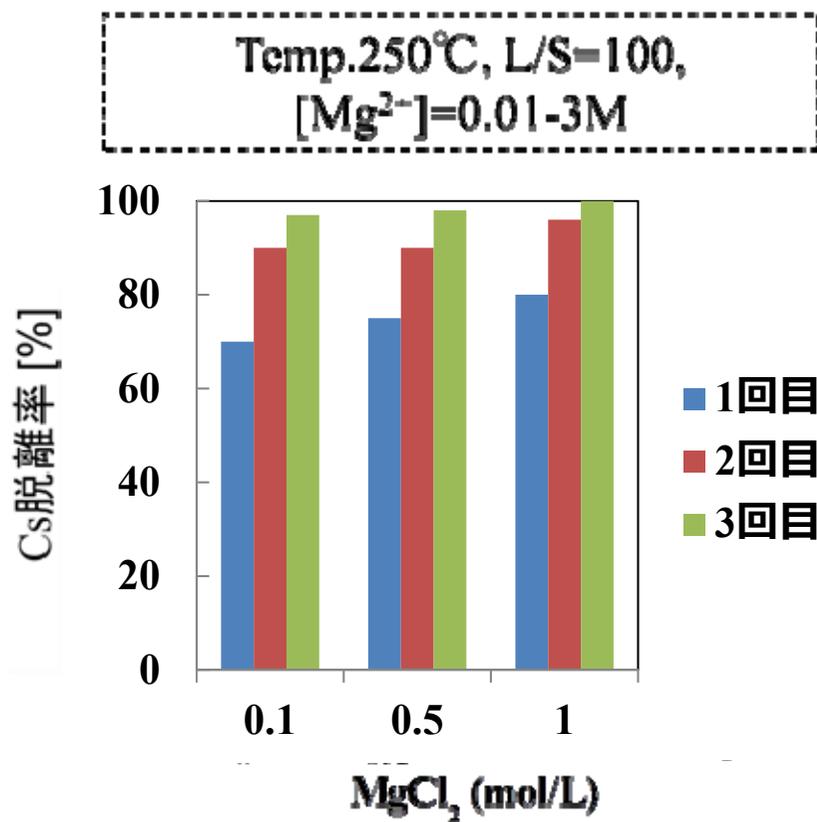
## (2) イオン含有亜臨界水によるCs脱離試験(ホット)

溶離液の $Mg^{2+}$ 濃度0.5Mにおいて2回の亜臨界水洗浄でほぼ90%、3回で96%のCs回収率を達成できた。本法により10万Bq/kg程度の汚染物であれば、十分に8000Bq/kg以下にできる見通しを得た。

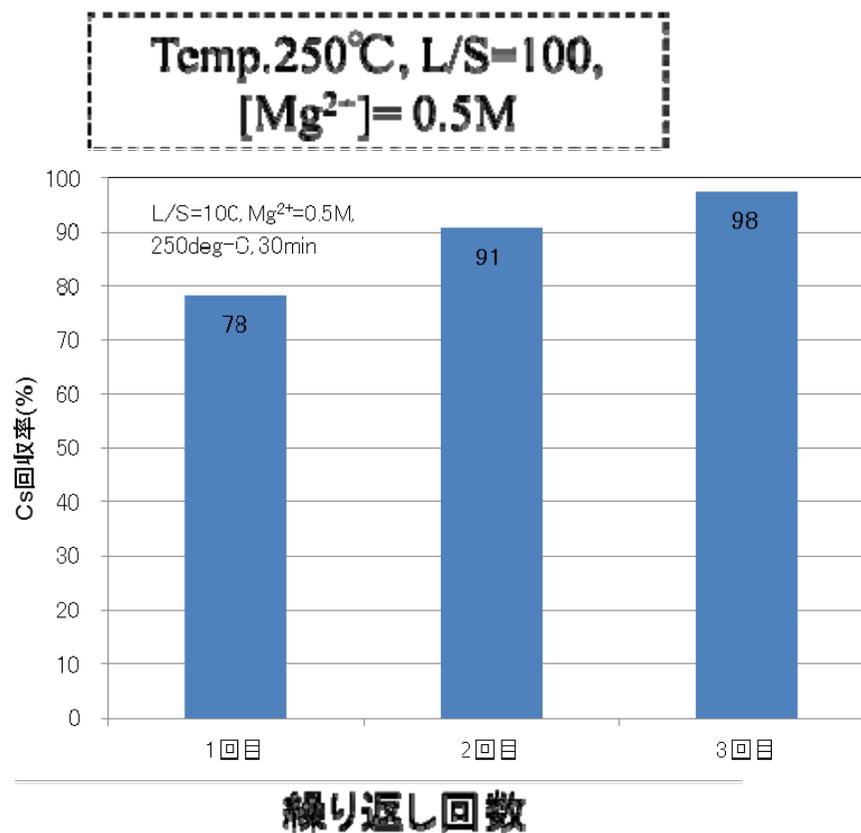


L/S=100で0.5 M  $Mg^{2+}$ 溶離液を用いた場合の繰り返し試験

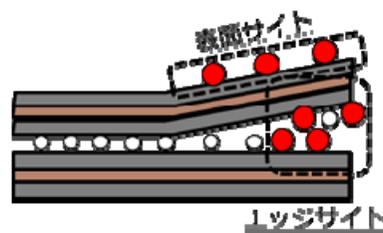
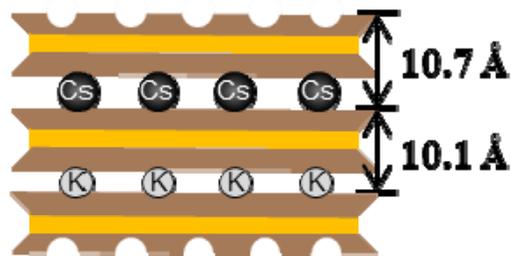
# Comparison between cold test and hot test (250°C)



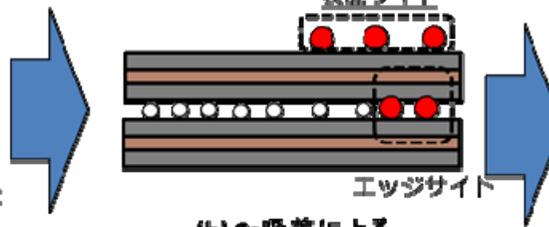
(a) コールド試験 (Mg<sup>2+</sup>濃度の影響)



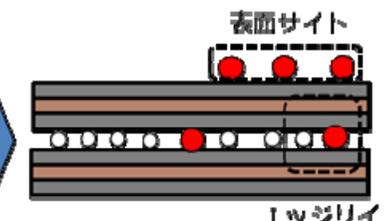
(b) ホット試験 (Mg<sup>2+</sup>濃度固定)



(a) Frayed Edge site へのCs吸着

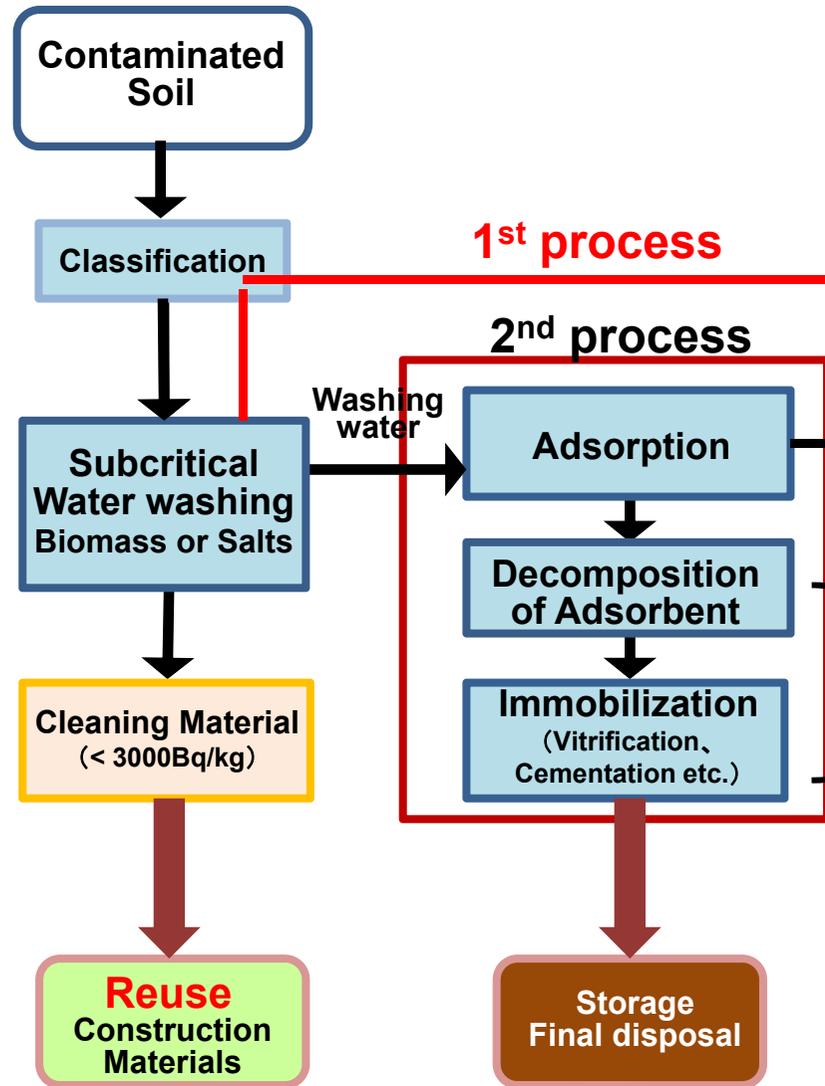


(b) Cs吸着による Edge部の修復



(c) 吸着したCsの拡散

# Volume Reduction Technology of Cs-contaminated Soil



In the 2<sup>nd</sup> process, Cs in washing water is concentrated by an adsorbent, ferric ferrocyanide and solidified into glass

## (1) Subcritical water washing

Release of Cs from Soil to water

## (2) Adsorption

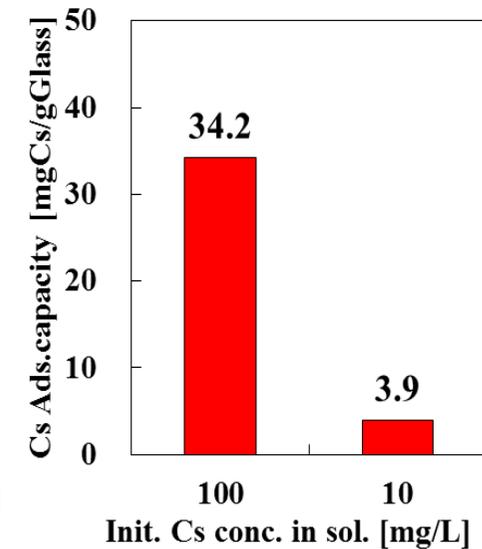
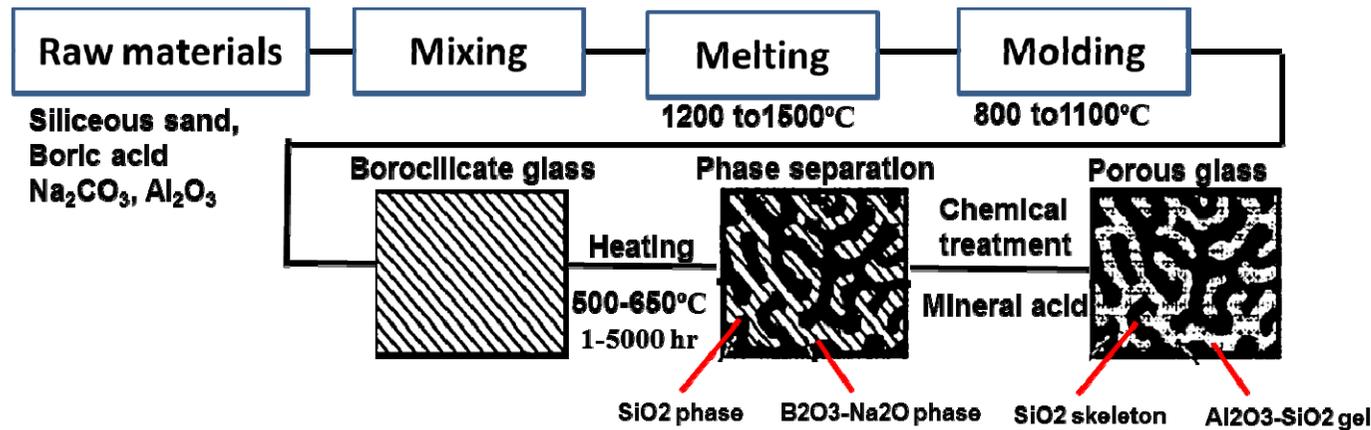
Selective recovery of Cs from washing water by ferric ferrocyanide

## (3) Solidification

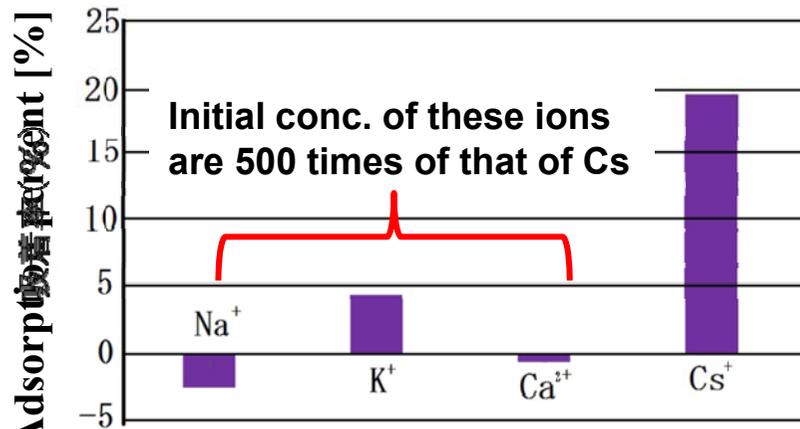
- Thermal Decomposition of ferric ferrocyanide
- Leaching of Cs from residue to water and diluted HNO<sub>3</sub>
- Immobilization of Cs in glass

# Future Adsorbent --- Porous Glass with Alminosilicate Gel

## 1. Synthesis of porous glass with aluminosilicate gel

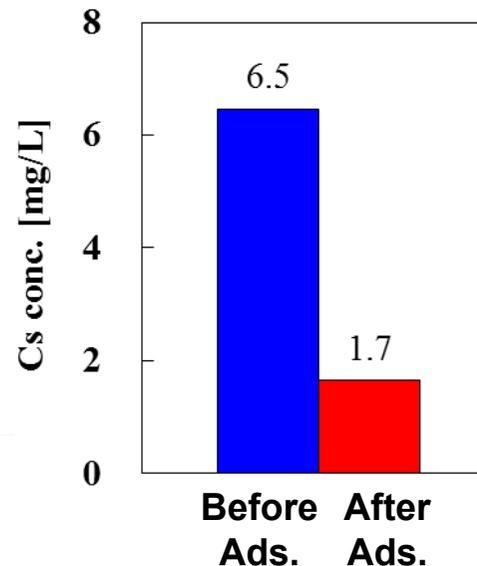


## 2. Selectivity



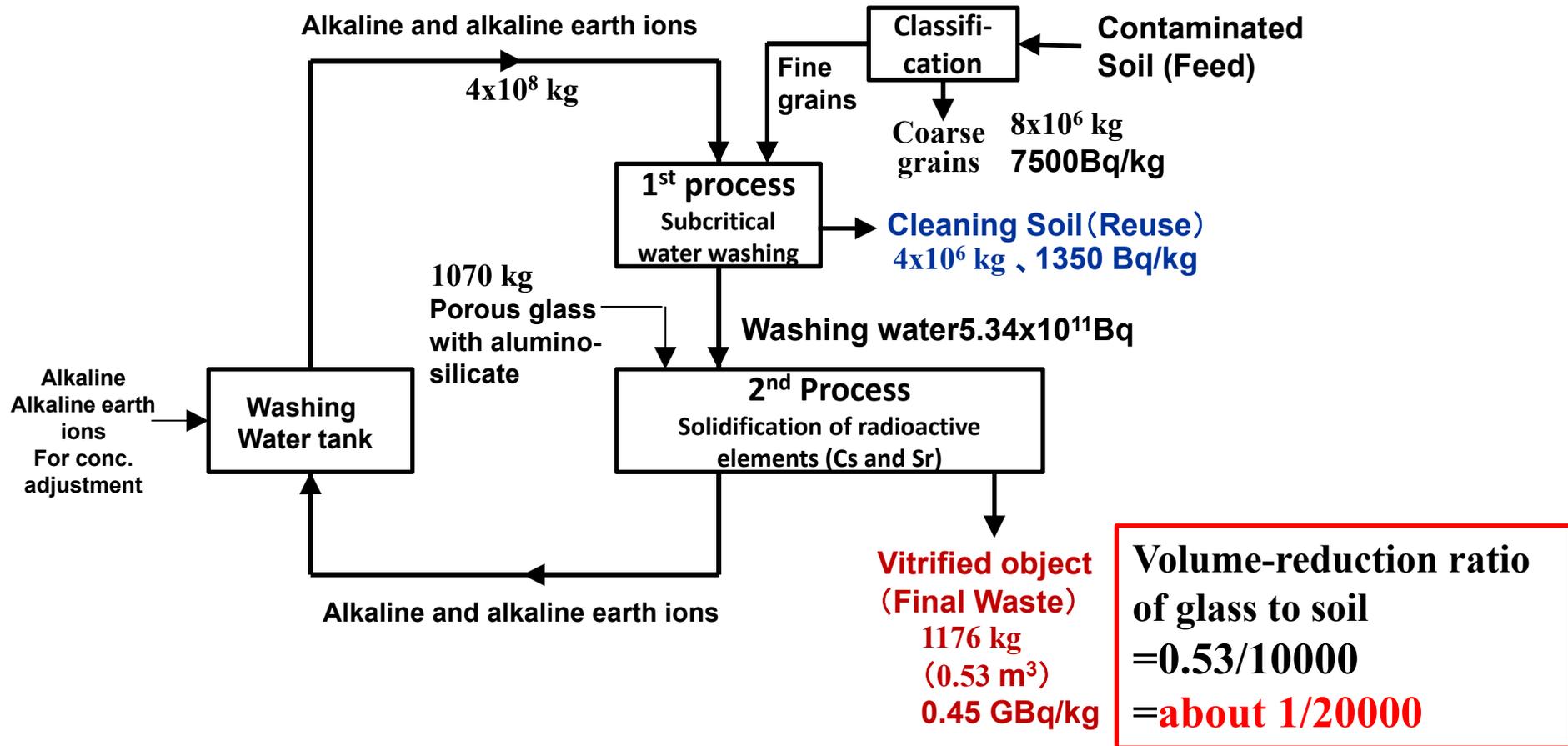
Cs is adsorbed selectively compared with Na, K and Ca.

## 3. Adsorption performance



- Washing solution after subcritical water washing was used for the Cs adsorption test.
- The content of aluminosilicate gel is 3 wt%. The concentration of Cs in the washing solution is 6.5 mg-Cs/L.
- 1g of the porous glass was added to 100mL of the washing solution.
- Cs was adsorbed selectively and the adsorption capacity is 0.6mg/g-Glass.

# Mass and Radioactive Balances of Future Treatment System



**Feed Conditions** (Continuous operation is assumed)

**Contaminated Soil**

**Feed :  $10,000 \text{ m}^3$  ( $1.2 \times 10^7$  kg),  
 Radioactivity : 50,000 Bq/kg**

**Operating conditions**

**Processing speed : 30 ~ 40 ton/day  
 Processing period : 300 days**

## Summary

- **A solid waste treatment system, which consists of two processes, subcritical water washing process and solidification process of radioactive elements, is proposed.**
- **Cs adsorbed strongly in clay minerals (vermiculite) was released to water effectively by the subcritical water washing process using organic acids (decomposition of clay minerals) or several salts (high-speed ion exchange ).**
- **In the case of the immobilization of Cs into glass, the high volume reduction of contaminated soil is attained by the introduction of the proposed system ( volume ratio of glass to soil = 1/20,000) .**

**Thank you for your attention**



**International collaborative Research Building**

# Thank you for your attention



## Contaminated soil

Photo: Yomiuri Shimbun

# Radioactive Wastes generated by Decommissioning of 1F NPP

- Gaseous waste, liquid waste and solid waste containing radioactive elements are generated by the decommissioning of Fukushima daiichi NPP.
- Gaseous and liquid wastes are immobilized in solid materials and the release of radioactive elements to the environment can be suppressed.

## Gaseous waste



exhaust  
gas



Radioactive elements in gaseous waste from reactor vessels is collected by filter

Solid waste

## Liquid waste



Radioactive elements in liquid waste is recovered in polluted water treatment facility



Solid waste

## Solid waste

Rubble etc.



Polluted metal, concrete, soil, cut tree, protective closing and filter

Secondary wastes from Treatment facility



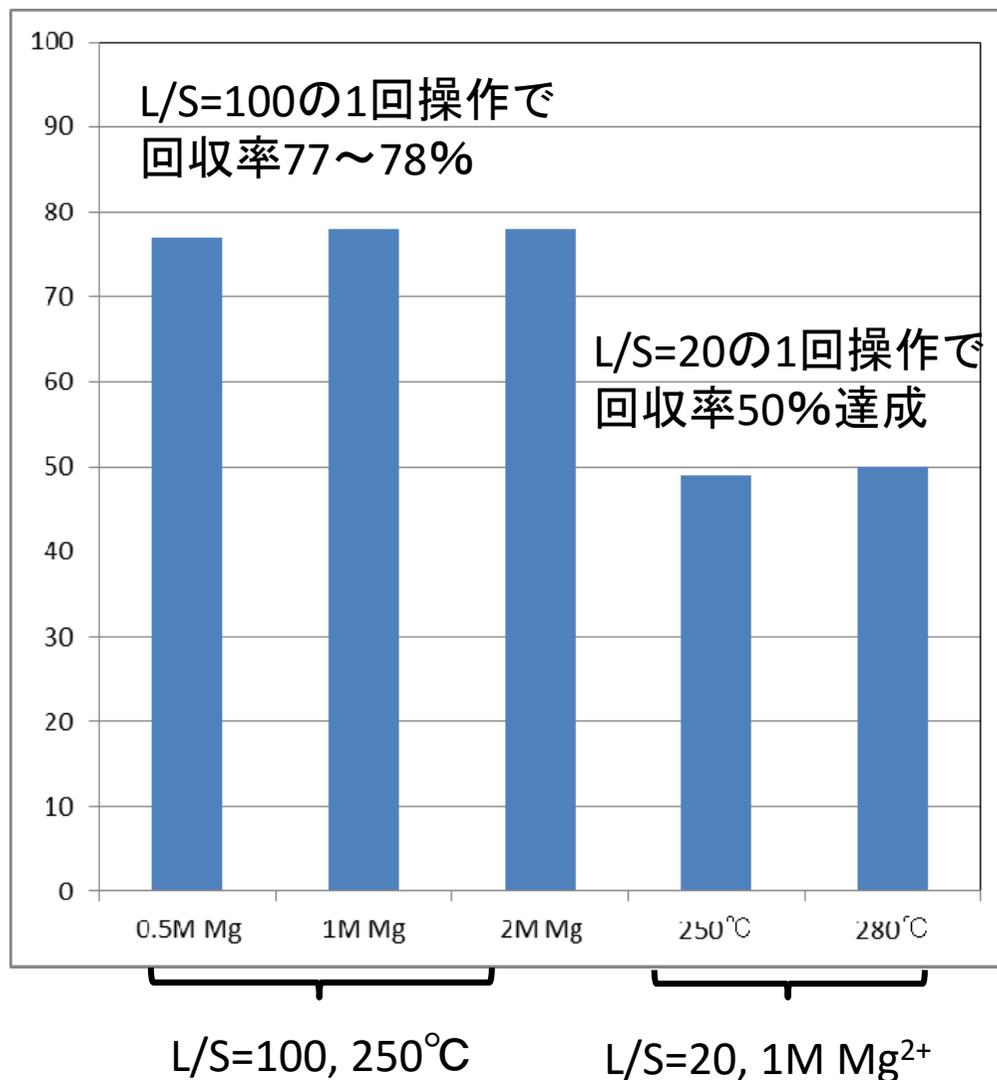
Spent Adsorbents, Sludge

Ref. 東京電力 福島第一原子力発電所 (1F)の廃止措置で発生する固体廃棄物の管理について (2016年4月)

	汚染土壌		非汚染土壌	
	分級前	分級後	分級前	分級後
SiO <sub>2</sub>	63.1	58.7	54.9	60.9
Al <sub>2</sub> O <sub>3</sub>	20.0	20.8	26.6	18.1
Fe <sub>2</sub> O <sub>3</sub>	11.5	12.7	12.0	11.1
TiO <sub>2</sub>	1.6	1.5	1.4	2.5
MnO	0.2	0.2	—	0.1
MgO	0.9	2.7	1.9	2.1
CaO	0.9	0.7	1.0	1.0
Na <sub>2</sub> O	0.3	0.9	0.4	1.5
K <sub>2</sub> O	1.2	1.5	1.4	1.8
P <sub>2</sub> O <sub>5</sub>	0.3	0.3	0.2	0.6
合計	100	100	100	100

分級前後 (<75 $\mu$ m) においても組成に大きな差異はない。

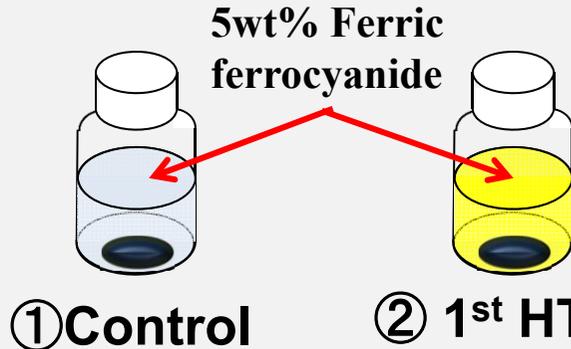
- ・L/S=20の場合、1回の処理で50%のCsを回収できることを確認
- ・L/S=20における繰り返し処理の効果を確認中



1回の亜臨界水洗浄におけるCs回収率の比較

# Adsorption Tests of Cs by Ferric Ferrocyanide

## Recovery of Cs released to hydrothermal water by ferric ferrocyanide



① Control

② 1<sup>st</sup> HTT

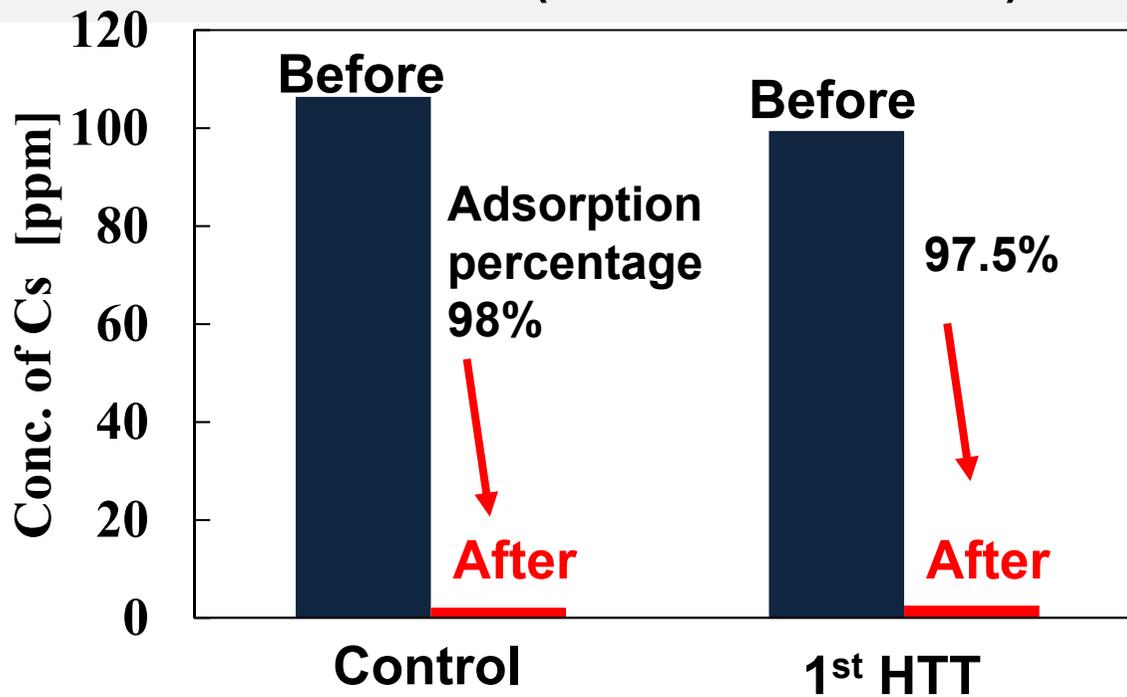
(Subcritical water washing)

① 100ppm CsCl solution (control)

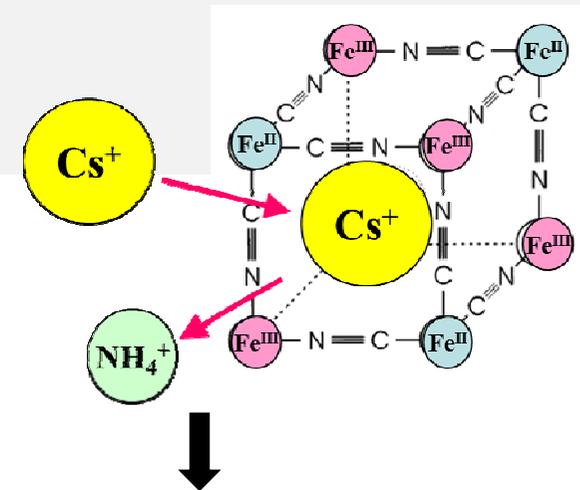
② Washing water after 1<sup>st</sup> HTT

- 5wt% Ferric ferrocyanide was added in these waters

- Stirred for 24h



Ads. amount 1.05 mg-Cs/g-PB    0.98 mg-Cs/g-PB



- Cs in hydrothermal water was adsorbed effectively by ferric ferrocyanide.

- Easy recovery of Cs by a conventional adsorbent

# Mass and Radioactive Balances of Proposed System

Adsorption of Cs  
(41g/Fe FC-kg)

Contaminated  
Soil

20,000 Bq/kg (total radioactivity  $2.4 \times 10^{11}$  Bq )  
Soil volume  $1 \times 10^4$  m<sup>3</sup> (  $1.2 \times 10^7$  kg)

(Conc. of <sup>133</sup>Cs 0.010g/Soil-kg)

Classification  
Process

Bulky soil  
 $8 \times 10^6$  kg  
(2/3 of total soil)  
3,000 Bq/kg ( $2.4 \times 10^{10}$  Bq)

Recycling Water

Fine grain  
 $4 \times 10^6$  kg  
(1/3 of total soil)  
54,000 Bq/kg ( $2.16 \times 10^{11}$  Bq)

Fe ferrocyanide  
adsorbent  
 $2.50 \times 10^3$  kg

Water Cleanup  
Process

Water  
 $4 \times 10^8$  kg

Organic Acid  
0.3 mol/L

Subcritical water  
Process  
(sep eff. 95wt%)

Washing water

$4 \times 10^8$  kg  
( $2.05 \times 10^{11}$  Bq  
 $\Rightarrow$  <sup>137</sup>Cs 0.064g)  
(<sup>133</sup>Cs  $\approx$  102600g)

Adsorption

Water  
 $4 \times 10^8$  kg

Residue  
 $4 \times 10^6$  kg  
2,700Bq/kg  
( $1.04 \times 10^{10}$  Bq)

Cs-adsorbent  $2.50 \times 10^3$  kg

Decomposition

Iron oxide  
 $1.65 \times 10^3$  kg

Volume Reduction =  $0.466 \text{ m}^3 / (1 \times 10^4 \text{ m}^3)$

$\approx 1/21442$

$\Rightarrow$  約 1/20000

<sup>137</sup>Cs + <sup>133</sup>Cs = 102600.319g  
(Recovery 100wt%)

Glass  
 $9.2 \times 10^2$  kg

Vitrification

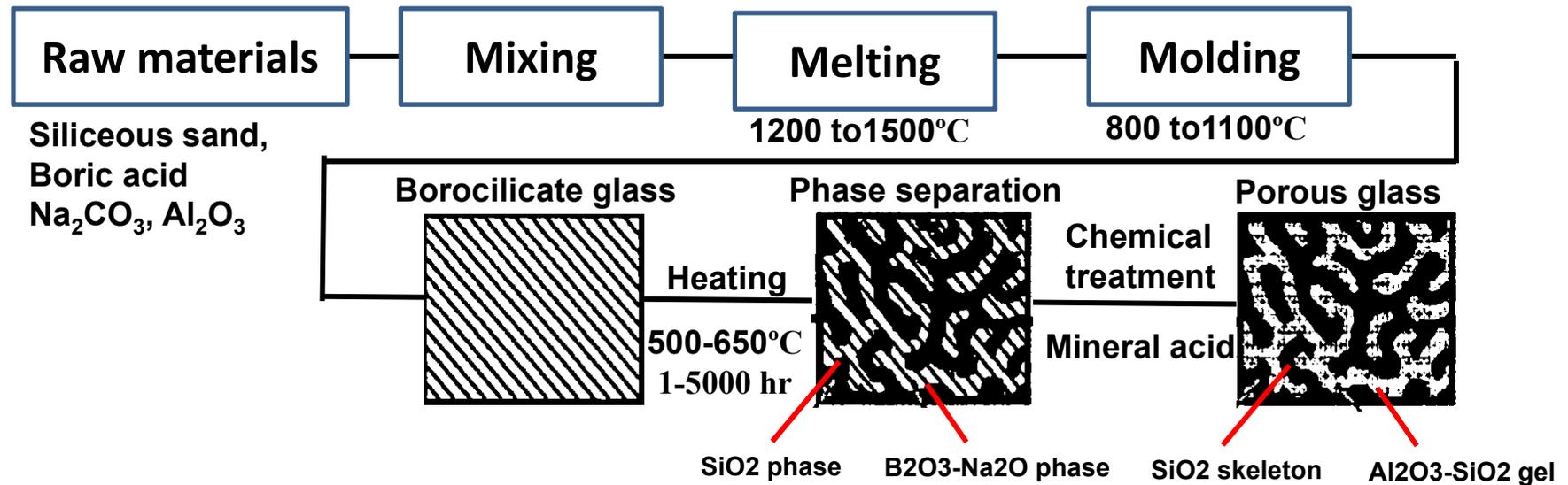
※ガラス固化体密度  
 $2.2 \times 10^3 \text{ kg/m}^3$  と仮定

Number of Glass Rods = 3  
(one Glass Rod = 170 L)

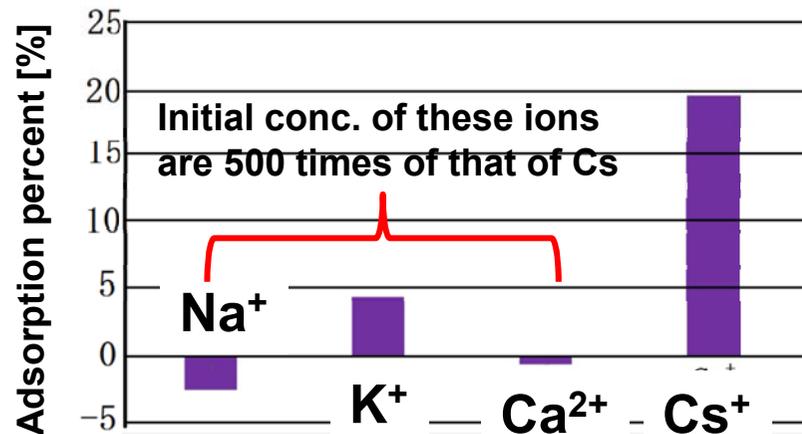
Vitrified Object (Cs:10wt%)  
 $1.03 \times 10^3$  kg ( $0.466 \text{ m}^3 = 466 \text{ L}$ )  
 $2.0 \times 10^8$  Bq/kg  
( $2.05 \times 10^{11}$  Bq)

# Adsorption of Cs by Porous Glass with Aluminosilicate Gel

## 1. Synthesis of porous glass with aluminosilicate gel



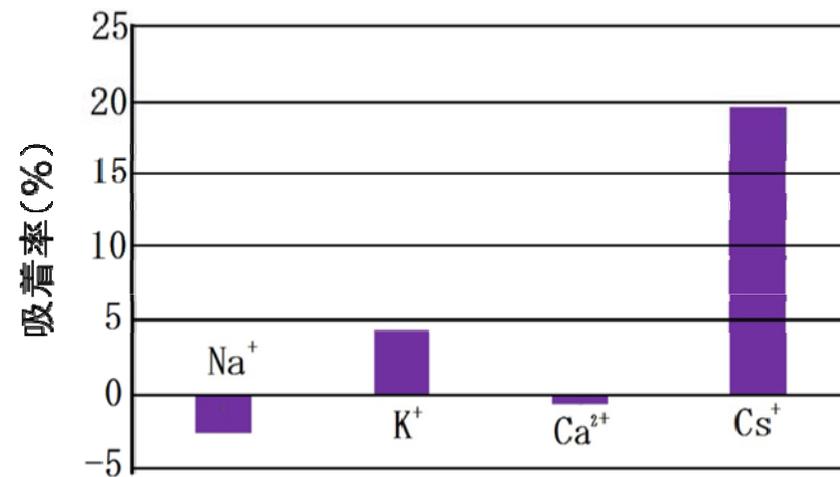
## 2. Metal selectivity



Cs is adsorbed selectively rather than Na, K and Cs

## 3. Adsorption amount

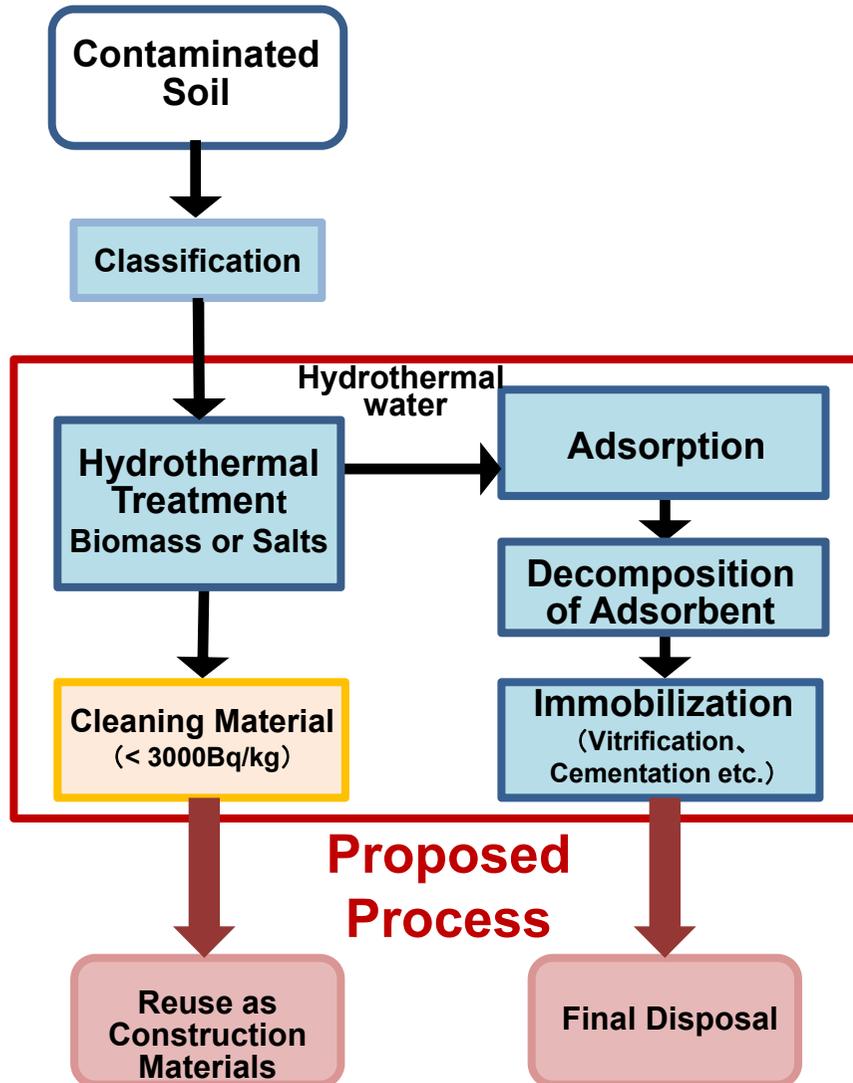
### アルミノシリケートゲル堆積型高ケイ酸ポーラスガラスのCs吸着特性



混合溶液中の各イオンに対する吸着率

図6 アルミノシリケートゲル堆積型高ケイ酸ポーラスガラスの吸着性能

# Volume Reduction Technology of Cs-contaminated Soil



A new volume-reduction process of Cs-contaminants is proposed for the secondary volume reduction during interim storage. This process has four steps as follow,

## (1) Hydrothermal

Release of Cs from Soil to water

## (2) Adsorption

Selective recovery of Cs from water

## (3) Decomposition

Cs Conc. by leaching from residue

## (4) Immobilization

Immobilization of Cs into glass

## Conventional Volume Reduction Methods of Contaminated Soil

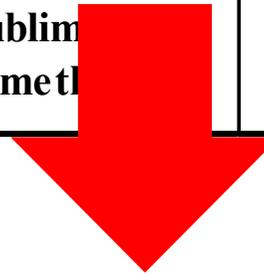
	Operator	Object	Volume Reduction Method	Temp. [°C]	Initial Radioactive Cs Conc. [Bq/kg]	Cs Removal [%]
Wet Process	Toshiba Co.	Soil	Oxalic acid washing	95	5,000	93% (for 5 times washing)
	Tokyo Tech.	Soil	Hydrothermal + Coagulant settling	200	55,000	75% (85% if adding blasting)
Dry Process	Taiheyo Cement Co.	Soil	sublimation method	1300	60,000	99%

This table shows conventional volume reduction technologies of contaminated soil. Toshiba process and Tokyo Tech process are classified as wet processes. Toshiba process is a typical washing process using oxalic acid.

Taiheyo cement process is classified as a dry process. This is a sublimation process. Cs in contaminated soil is transferred to gaseous phase.

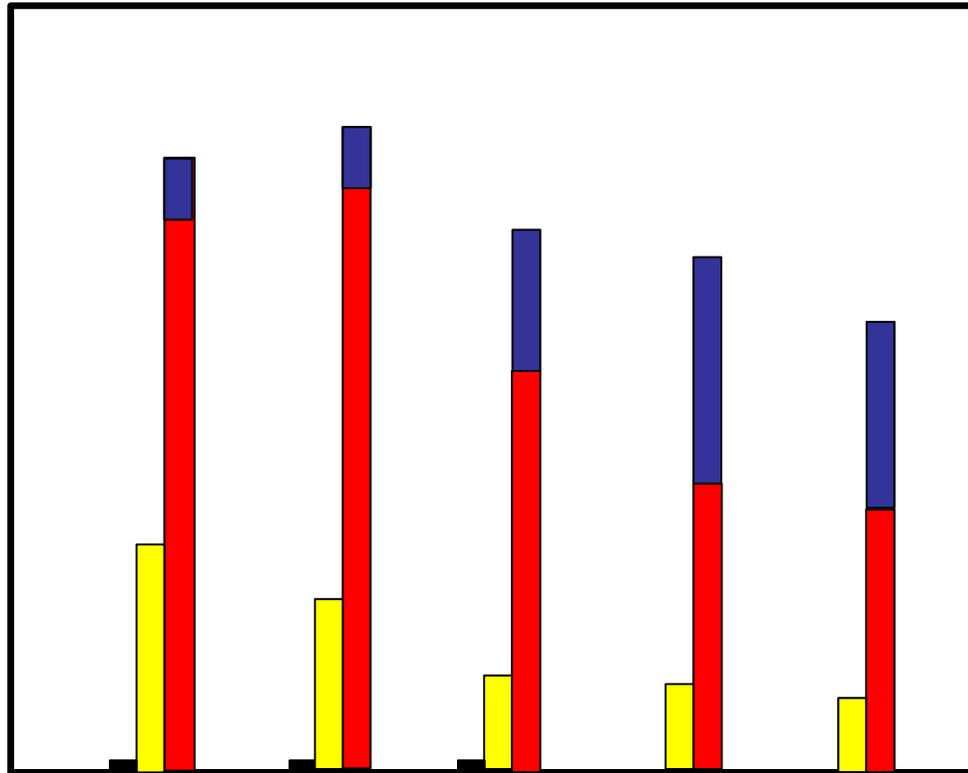
## Conventional Volume Reduction Methods of Contaminated Soil

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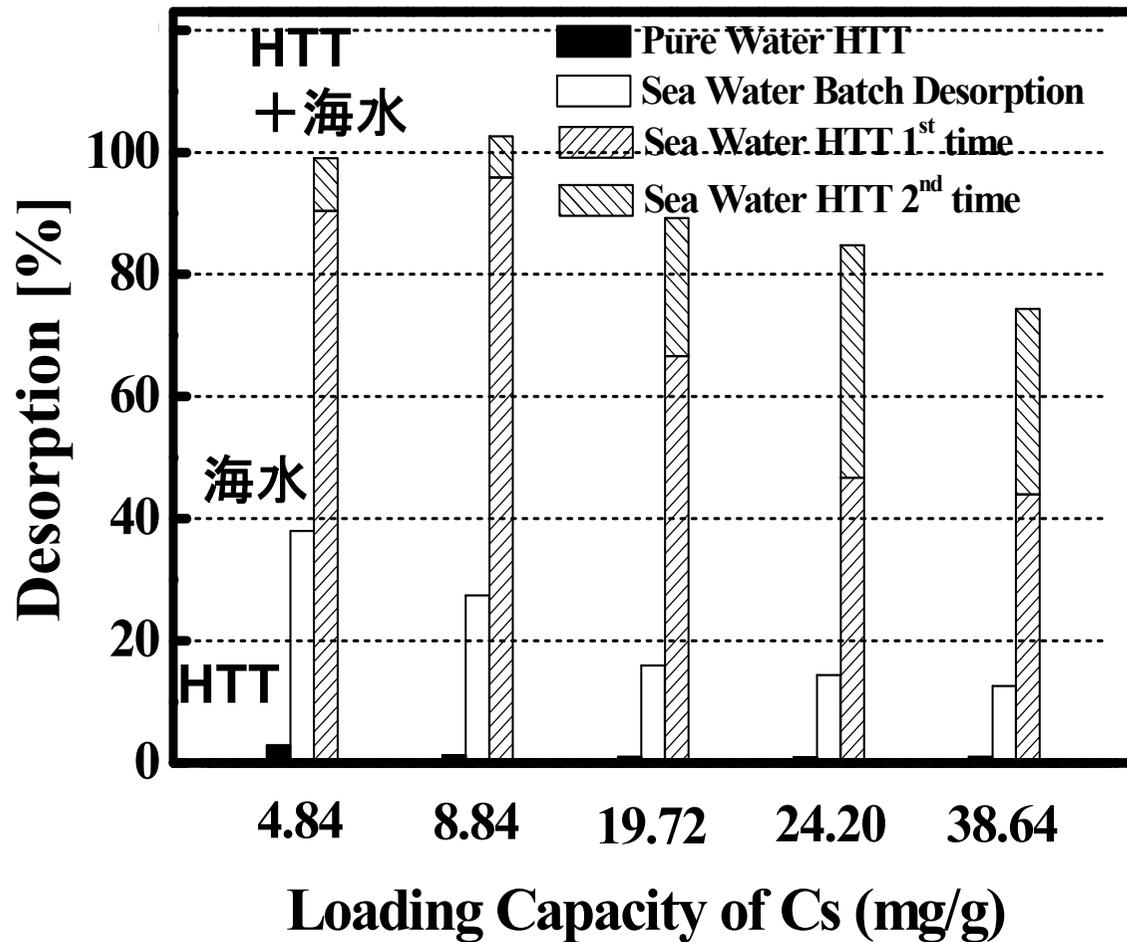
**Recent Progress of Hydrothermal process for the volume reduction of contaminated soil**

# HTT Desorption



- Pure water is no use to desorb Cs
- Sea water HTT significantly increases the desorption.

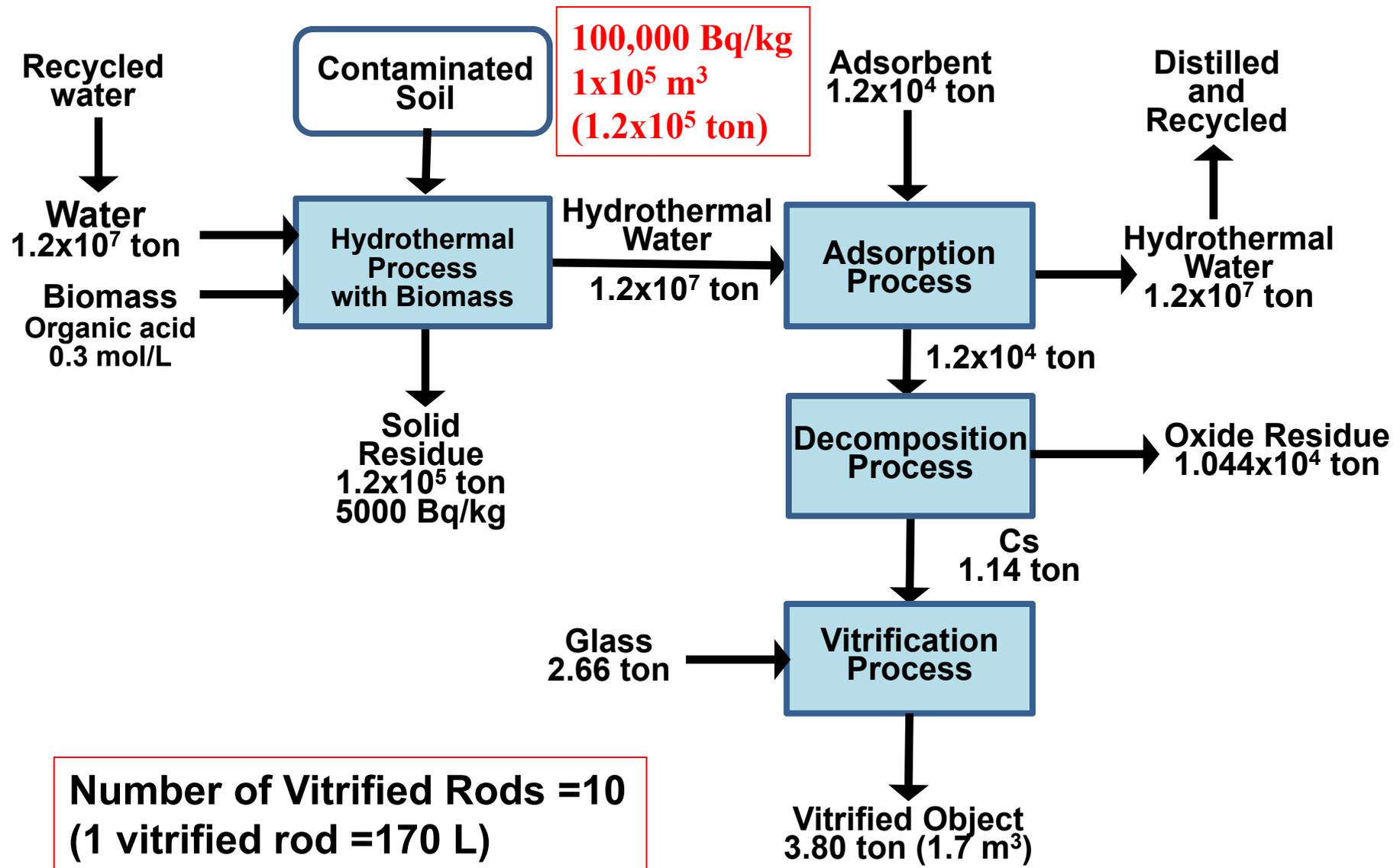
# バーミキュライトのCsイオン交換性能の比較



V/m =  
50ml/0.5g;  
250°C, t: 30 min

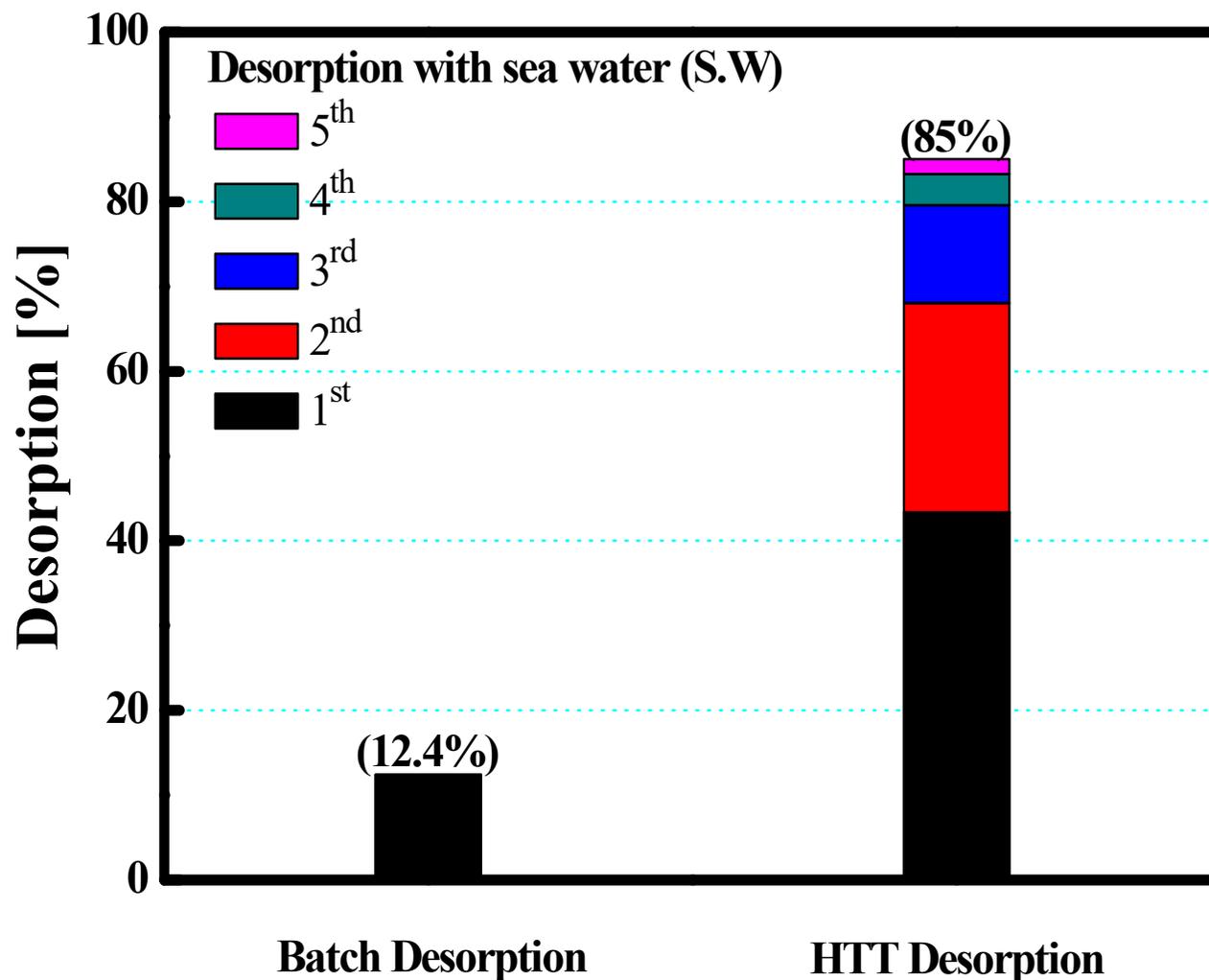
- Pure water is not useful for the desorption of Cs
- Sea water+HTT promotes the desorption of Cs.

# Mass Balance of Proposed Volume-reduction Process



**Volume Reduction =  $1.7 \times 10^{-5} = 0.0017\%$**

# Sequential Desorption using Sea Water

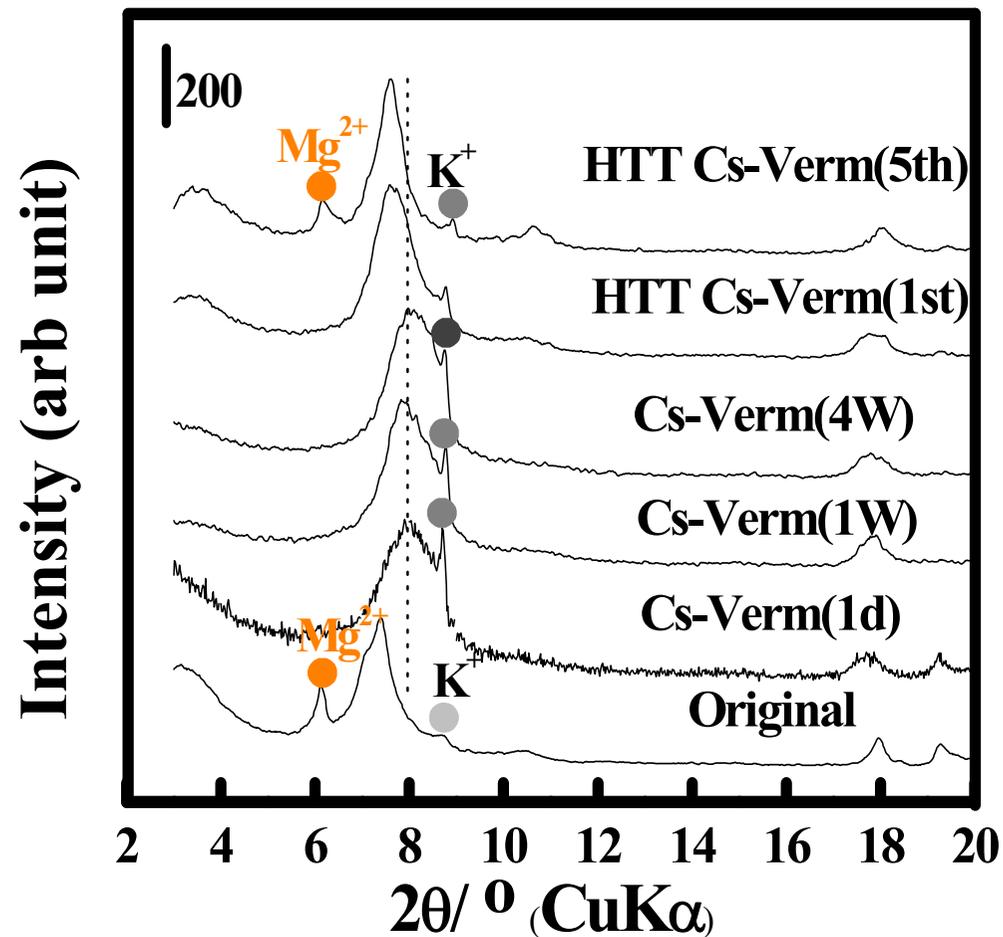


Adsorption: 28d; 10  
1000ppr  
HTT : V/m: 50ml/0.5g

常温脱着

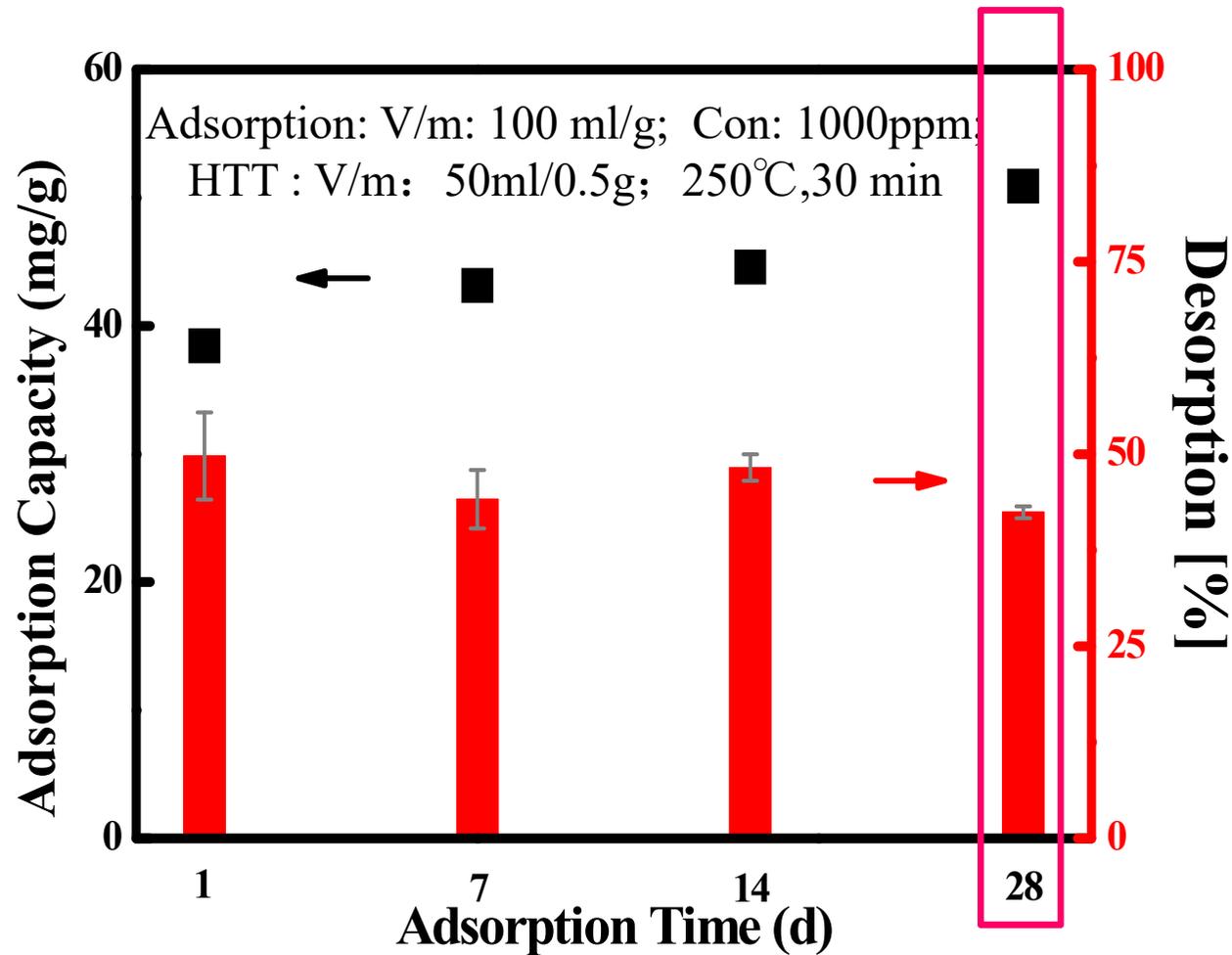
HTT脱着

## XRD Results With Sea Water Desorption



- The peak of Mg<sup>2+</sup> **disappears** after Cs<sup>+</sup> adsorption.
- The peak of Mg<sup>2+</sup> **appears** after HTT desorption again.

## Cs吸着後のエージング時間の影響



- Adsorption is **increased** with the time.
- Desorption is **decreased** with the time.

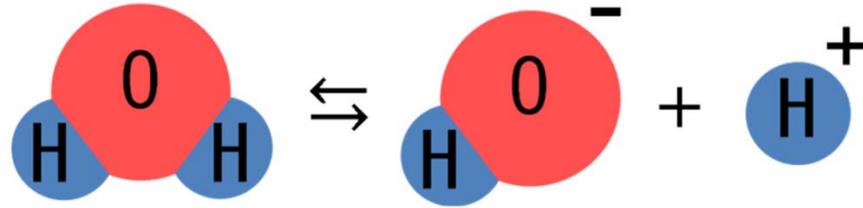
# Fukushima EcoTech

Fukushima EcoTech is located at border between Tomioka and Naraha. The leachate-controlled landfill site is on the top surface of the hardly-permeable ground. Double water barrier sheet is laid on the ground.



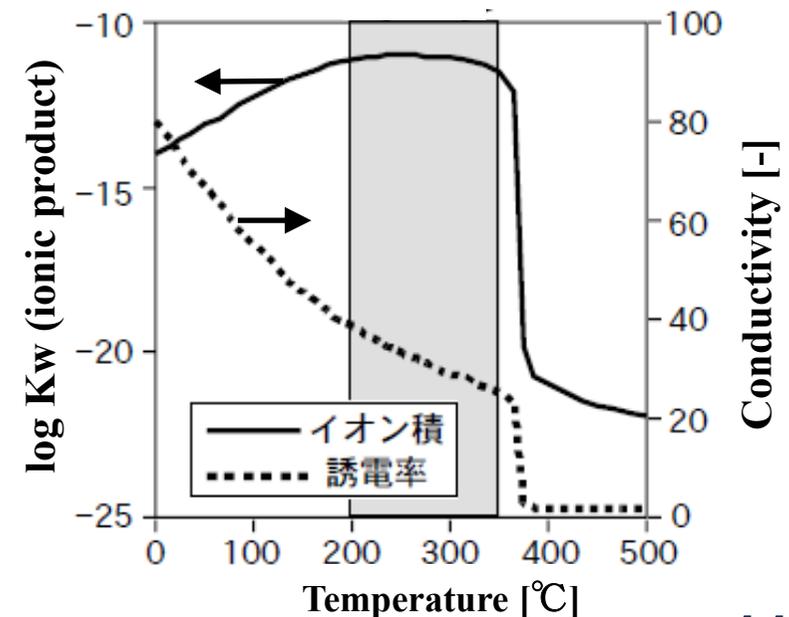
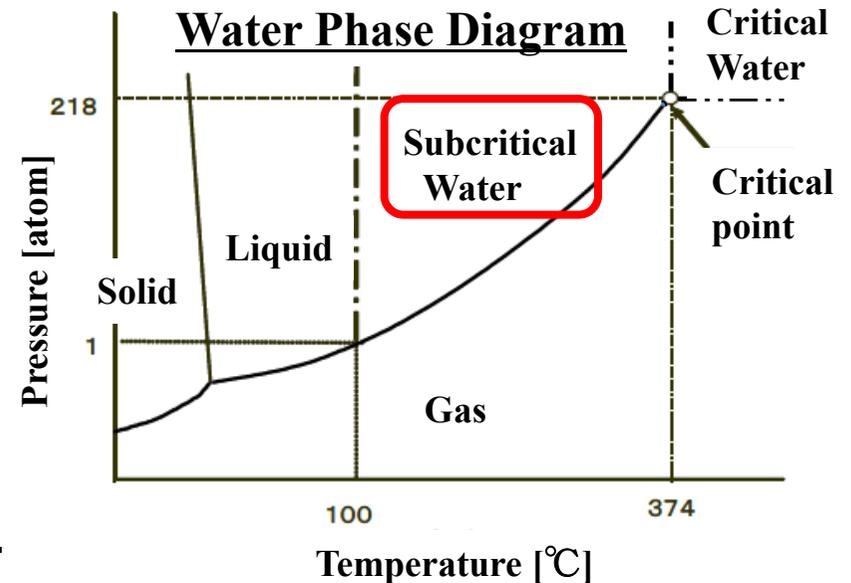
# Hydrothermal Decomposition using Subcritical Water

Promoting water solubilization of polluted materials by hydrothermal decomposition using subcritical water



Ionic product of water becomes largest in subcritical state under high temperature (240-280°C) and high pressure (3-4 MPa).

- ① Cs adsorbed in organic materials is recovered to water phase by the high decomposition effect of organic materials.
- ② Cs adsorbed in inorganic materials is recovered to the aqueous phase by the high-speed ion exchange effect.
- ③ Only water is used in the decomposition process. No secondary waste is formed.



Ionic Product and Conductivity of Water

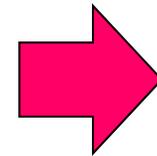
# Hydrothermal Decomposition of Contaminated Plants

## Hydrothermal Operation

- Water is added twice the sample
- 1. Hydrothermal operation at 200°C (14 atm) for 30 min
- 2. Gas venting at 170°C
- 3. Sampling at 70°C



**Before**



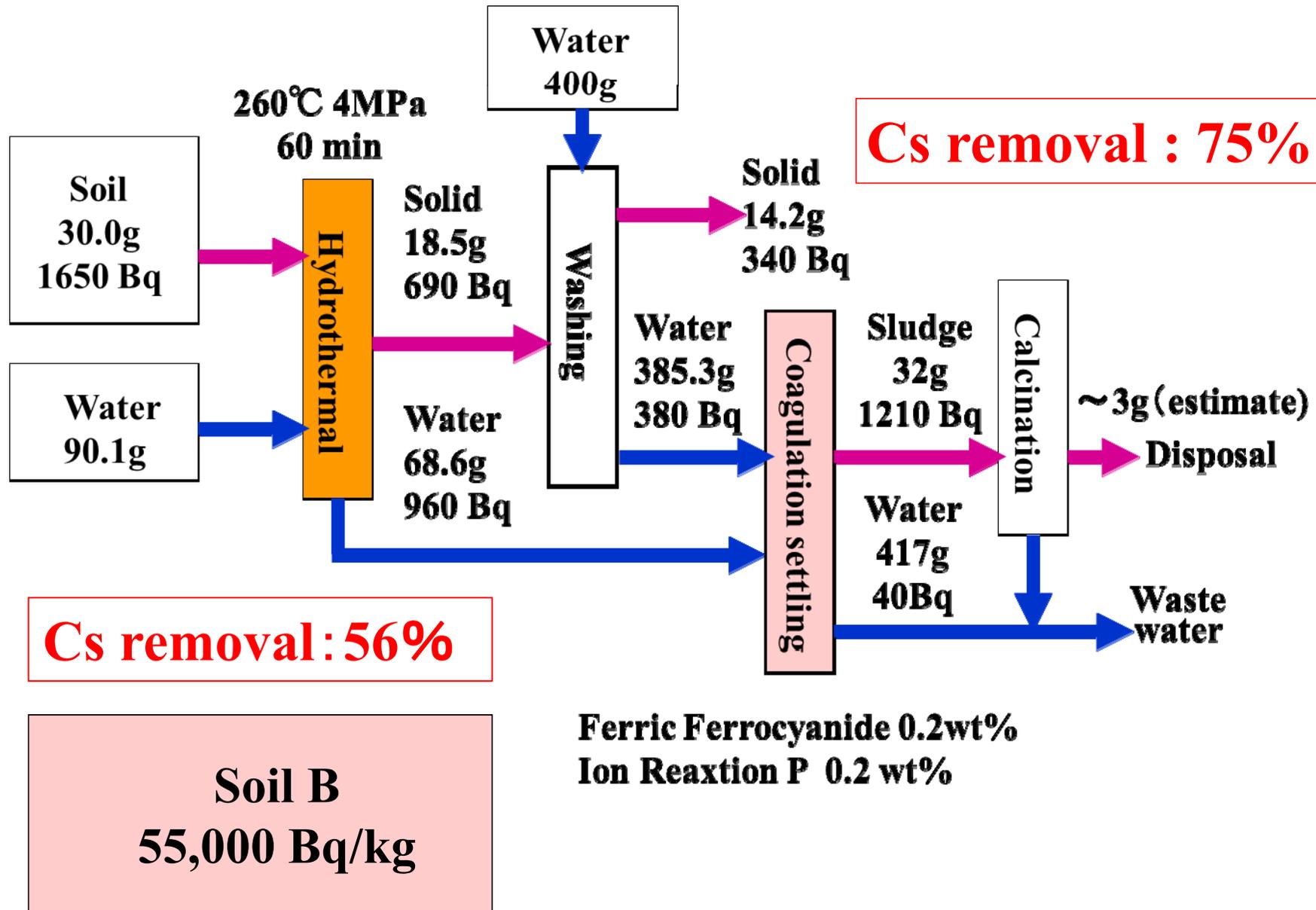
**After**



Organic compounds were decomposed by the hydrothermal operation. Most of solid components were decomposed to light-weight compounds, which were dissolved to water phase. Cs is also transferred to water phase.

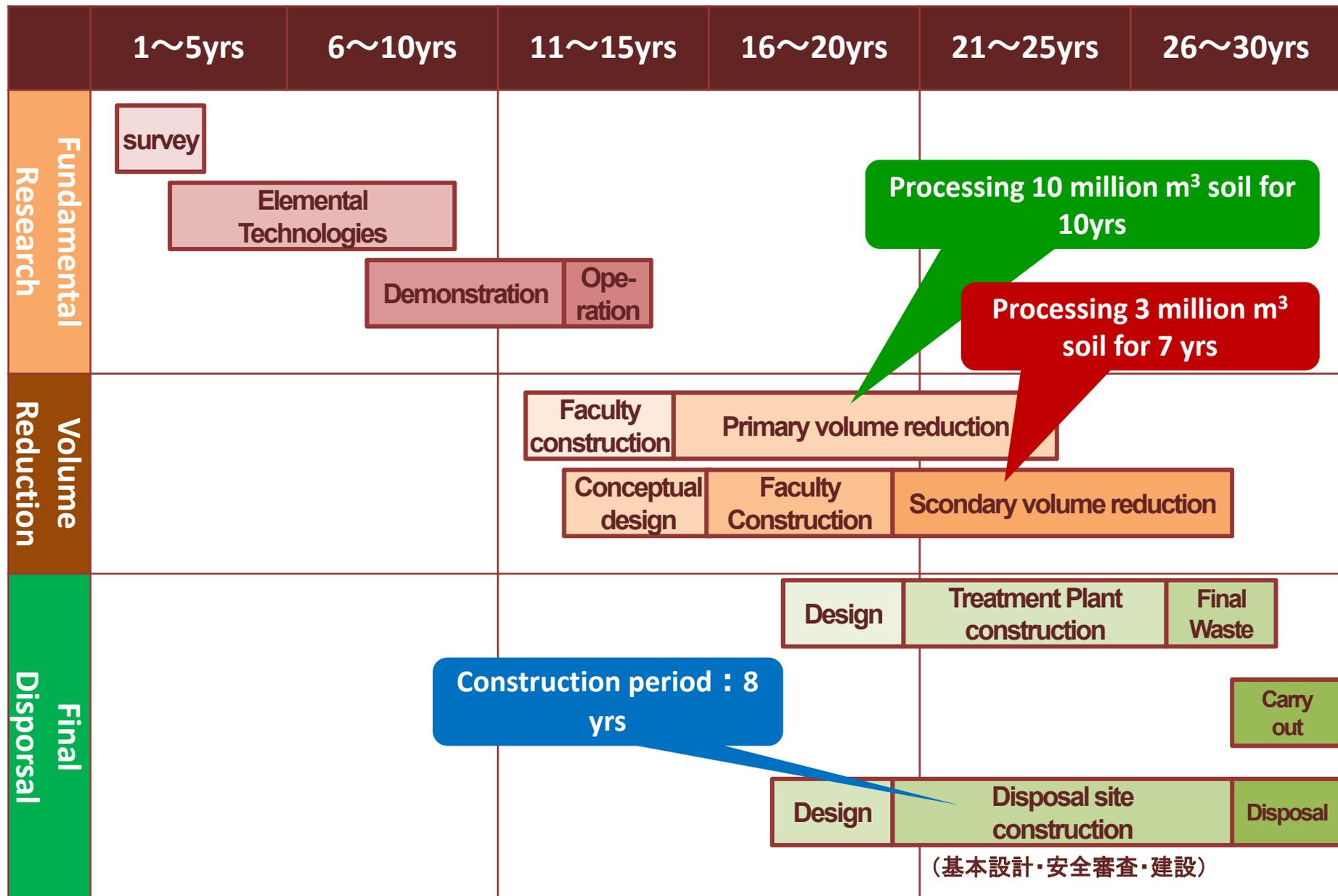


# Removal Test of Cs from Soil B



# Time Schedule of Volume-reduction of Contaminated Soil

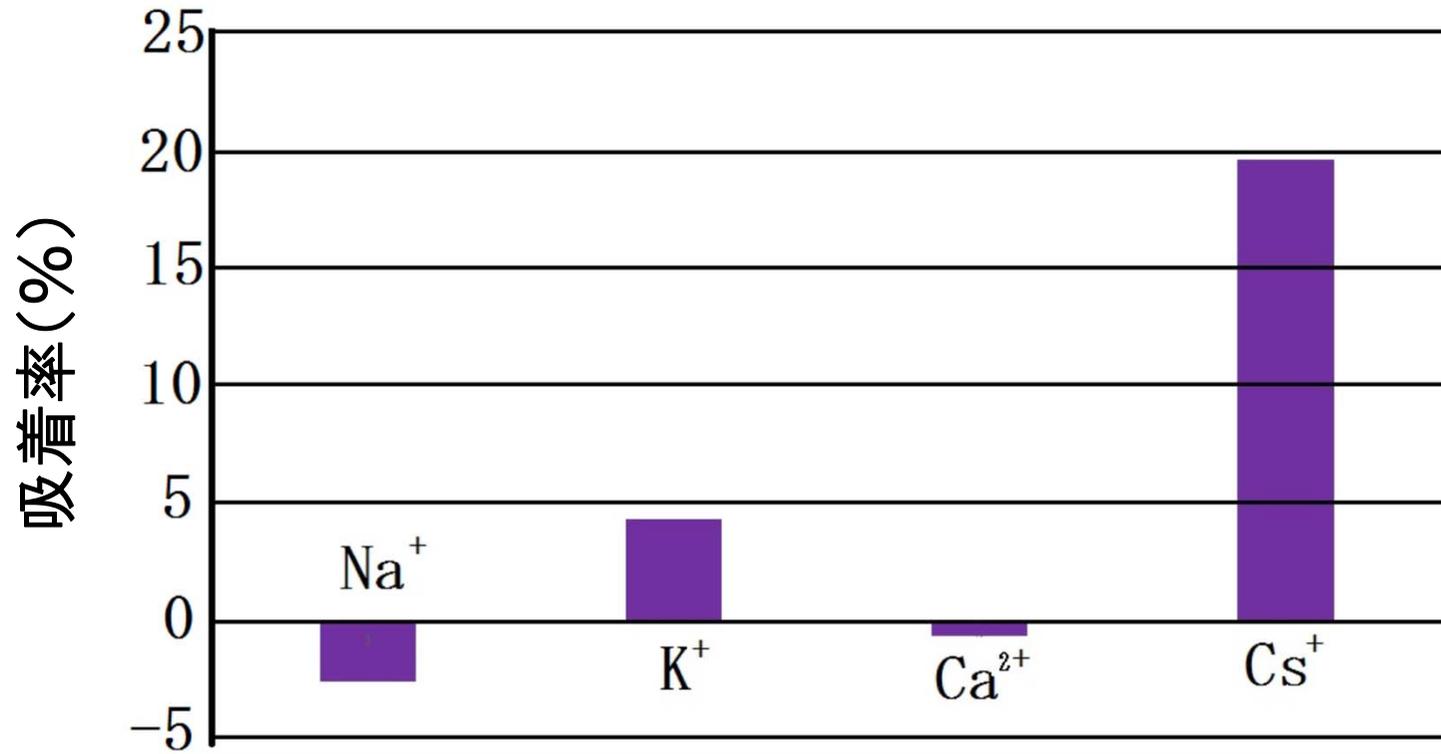
Proposed by RANDEC (Radwaste and Decommissioning Center)



## Generation of Contaminants in Fukushima Prefecture

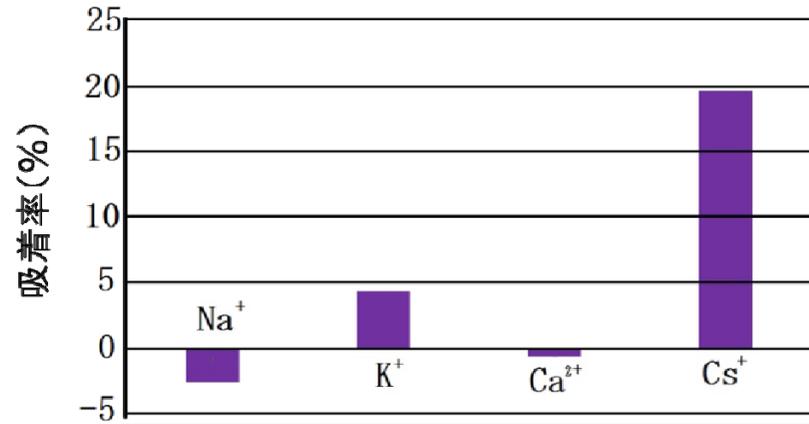
<b>Contaminants</b>	<b>Volume [m<sup>3</sup>]</b>
<b>Contaminated Soil below 8,000 Bq/kg</b>	<b>10,060,000</b>
<b>Contaminated Soil from 8,000 Bq/kg to 100,000 Bq/kg</b>	<b>10,350,000</b>
<b>Contaminated Soil above 100,000 Bq/kg</b>	<b>10,000</b>
<b>Ash of Combustible Decontamination Wastes</b>	<b>1,550,000</b>
<b>Decontamination Wastes above 100,000 Bq/kg</b>	<b>20,000</b>

Cs+1 ppmとCs+に対しモル比濃度500倍のNa+、K+、Ca2+混合  
溶液に対して示す各イオンに対する吸着率



混合溶液中の各イオンに対する吸着率

Cs+1 ppmとCs+に対しモル比濃度500倍のNa+、K+、Ca2+混合溶液に対して示す各イオンに対する吸着率



混合溶液中の各イオンに対する吸着率

図3 アルミナゲル堆積型高珪酸ポーラスガラスの吸着性能